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1 Evaluation of lethal control of introduced sugar gliders as a tool to relieve bird nest predation

2 **Running Head:** Culling an introduced bird predator

3 **Abstract**

4 Lethal control of invasive mammalian predators can be controversial and is rarely a 'silver bullet' for
5 conservation problems. Evaluating the efficacy of lethal control is important for demonstrating the
6 benefits to threatened species are real and detecting unexpected perverse outcomes. We
7 implemented a pilot study to evaluate if lethal control of introduced sugar gliders *Petaurus breviceps*
8 can reduce the rate of nest predation on Tasmanian hollow nesting birds including swift parrots
9 (*Lathamus discolor*). Using a before-after-control-impact design, we implemented a lethal control
10 treatment whereby we attempted to remove sugar gliders from three treatment sites. In each time
11 period across sites we monitored quail eggs in nest boxes to record predation, and used cameras to
12 detect sugar gliders. We caught nine sugar gliders over three treatment sites. The model best
13 supported by the data indicated an effect of site × time period on both egg survival and the rate of
14 glider detection on cameras. There was no support for an effect of treatment on our data. We also
15 recorded predation of a real swift parrot nest by sugar gliders at a treatment site where we recorded
16 no predation of quail eggs. Our pilot study shows that at small scales, intensive lethal control of
17 gliders yields low capture rates and no discernible effect on the metrics we measured. We conclude
18 that alternative approaches to controlling the impact of sugar gliders, such as habitat protection, are
19 critical in this study system before lethal control is widely implemented as a management tool.

20 **Key Words:** common starling *Sturnus vulgaris*, conservation management, nest survival, predator
21 control, sugar glider *Petaurus breviceps*, swift parrot *Lathamus discolor*

22 **Introduction**

23 Invasive predatory mammals are a major threat to global biodiversity (Woinarski et al., 2015,
24 Medina et al., 2011, Szabo et al., 2012), and alien predators can have double the impact of native

25 predators (Salo et al., 2007). A common approach used by land managers to relieve the negative
26 impacts on threatened prey species from invasive mammals is lethal control (Doherty and Ritchie,
27 2017), and similar approaches are used in agricultural contexts as well (Van Eeden et al., 2018).
28 However, a review of the evidence from lethal control programs (Doherty and Ritchie, 2017)
29 identified four common perverse outcomes of lethal predator control: herbivore and mesopredator
30 release, disrupted predator social systems, predator immigration and ethical concerns. Doherty and
31 Ritchie (2017) recommend that adaptive, evidence-based approaches should be used in the
32 implementation of lethal control of invasive predators so that the efficacy of interventions can be
33 evaluated adequately. Global meta analysis have shown that the main determinant of management
34 success is the efficiency of the approach used to manipulate predator populations (Salo et al., 2010).
35 Here we report the results of a pilot study in Tasmania, Australia, aimed at relieving predation on
36 hollow nesting birds by sugar gliders *Petaurus breviceps* (Stojanovic et al., 2014). Sugar gliders are
37 introduced to Tasmania (Campbell et al., 2015), and are a severe threat to nesting swift parrots
38 *Lathamus discolor* (Heinsohn et al., 2015). There is intense conservation interest in relieving
39 predation pressure on swift parrots because they are critically endangered (Heinsohn et al., 2015).
40 Non-lethal techniques like predator exclusion devices on nest boxes (Stojanovic et al., 2019) offer
41 protection to individual nests, but finding a way of protecting nests in natural tree hollows over
42 larger spatial scales is crucial because only a fraction of parrots utilize nest boxes. Trials of predatory
43 owl call broadcasts at night did not reduce sugar glider predation on real or artificial bird nests
44 (Owens et al., 2020). These mixed results do not provide effective long-term, large-scale tools for
45 protecting swift parrots from sugar gliders.

46 Recently, Natural Resource Management South implemented a pilot study to evaluate the efficacy of
47 lethal control as an alternative approach to reducing predation on nesting birds. Using a BACI design,
48 we aimed to identify whether lethal control of sugar gliders: (i) reduced predation on artificial nests,

49 (ii) reduced sugar glider detection rates on remote cameras, and (iii) is logistically feasible at the
50 spatial scales necessary to protect swift parrots.

51 **Methods**

52 *Study sites and treatment groups*

53 The pilot study took place at six locations in Tasmania, Australia, where swift parrots and sugar
54 gliders are known to be sympatric (Heinsohn et al., 2015). The control sites were at Buckland (lat:
55 42°31', long: 147°39'), Lake Leake (lat:42°1', long: 147°49') and the Meehan Range (lat: 42°49', long:
56 147°24'). The treatment sites were at Tooms Lake (lat:42°13', long: 147°46'), Rheban (42°36', long:
57 147°54') and Southport Lagoon (lat: 43°29', long: 146°55'). All sites were dominated by dry
58 woodland, and understory composition ranged from grassy to shrubby.

59 The treatment involved two capture techniques: (i) active trapping using Mawbey traps, and (ii)
60 passive trapping using nest boxes fitted with doors operable by a person on the ground. A
61 veterinarian euthanized trapped sugar gliders by using lethal injection. The location of both trap
62 types was haphazard within swift parrot nesting habitat, and all sites had at least one natural swift
63 parrot nest hollow within the array of nest boxes and traps. Our study involved a 'before' period (14
64 nights duration) when all sites were established and 20 nest boxes, 10 Mawbey traps (deactivated)
65 and five camera traps baited with universal mammal bait were deployed. The 'after' period at
66 treatment sites involved three trapping sessions of four nights duration each. Due to logistic
67 constraints, we deployed personnel at only one site in a given week, so there was an interval of two
68 weeks between trapping sessions at a given site. After the completion of the nine week 'after'
69 period, we ceased trapping at sites for three months. We then implemented a 'long after' period,
70 comprising a further two days of monitoring at each treatment site, with an interval of 21 days
71 between egg deployment and follow up monitoring.

72 We assessed the impact of our treatment in two ways. At both treatment and control sites we
73 deployed one quail egg in each of the 20 nest boxes at the start of each time period. At the end of
74 each time period we checked whether quail eggs had survived the interval. We monitored glider
75 occurrence at the sites using cameras, and tallied the total number of nights in each time period at
76 all sites that we detected sugar gliders.

77 We spaced nest boxes and cameras ~30m apart within a given site to make monitoring feasible. This
78 reflects the natural distribution of natural nests of swift parrots, but it is likely that the same
79 individual sugar gliders preyed on multiple nests at the site level, and so we include site as a fixed
80 effect in our analysis below. However, we assumed that predation events at a given nest box over
81 successive time periods are independent of one another. Sugar gliders are the main predator of bird
82 nests in the nest boxes we used (Stojanovic et al., 2019), but we also recorded high occupancy of
83 nest boxes by common starlings *Sturnus vulgaris* at Lake Leake and Buckland. Common starlings
84 compete with swift parrot for nests and destroy their eggs (D.S. unpublished data). We included data
85 from nest boxes where common starlings destroyed the quail eggs because sugar glider removal
86 may result in increased common starling abundance via relaxed predation pressure at artificially
87 abundant nesting sites, creating the potential for a perverse outcome for parrots.

88 *Analysis*

89 We compared competing models using Akaike's Information Criterion, but corrected for small
90 sample sizes (AICc), and we considered that models within 2 Δ AICc had equivalent support (Burnham
91 and Anderson, 2002). All analyses were undertaken in R (R Core Team, 2017). We used predation of
92 quail eggs as a response variable, and fitted generalized linear models site ID, treatment, time
93 period, site ID \times time period and treatment \times time period as fixed effects. We used a binomial error
94 distribution, and to account for differences in the duration of time periods between some sites we
95 included the number of nights duration of each period as an offset term in the models. We never
96 observed predation of quail eggs at Southport, so we excluded this site from analysis. We also used

97 the number of nights sugar gliders were detected on cameras as a response variable in generalized
98 linear models and included the same fixed effects as above. We used a Poisson error distribution and
99 again included an offset term to account for the duration of each time period. We recorded no sugar
100 gliders in the first two time periods at Buckland, so we excluded this site from analysis.

101 **Results**

102 Nine sugar gliders were captured during the implementation of the trapping treatment at Rheban
103 (n= 3), Tooms Lake (n = 5) and Southport (n = 1).

104 The model best supported by the data for quail egg survival included the interaction between site
105 and time (Table 1). Modeled estimates of predation probability are presented with confidence
106 intervals in Figure 1. The very wide overlapping confidence limits indicate that there was low
107 confidence in modeled estimates at some sites, and high between-site variation in predation rates
108 over the study. The low support for the treatment × time period model ($\Delta AICc > 2$; Table 1) allows us
109 to reject the sugar glider lethal control treatment as a potential explanation for the rate of predation
110 on quail eggs. Instead, the model best supported by the data indicated that site level factors were
111 the best predictor of quail egg survival. For example, high predation rates at Buckland and Lake
112 Leake were largely attributable to common starling occupancy of nest boxes. Removal of common
113 starling nests did not reduce predation of quail eggs because in the interval between nest box
114 checks, common starlings were able to rebuild their nests and replace their clutches (as well as
115 remove quail eggs from boxes). One treatment site (Southport) had no predation on quail eggs, but
116 a real swift parrot nest within the study site failed due to sugar glider predation in the midst of the
117 trapping effort. The other treatment sites had relatively comparable predation rates to controls,
118 driven by sugar gliders throughout the trapping period, with a possible increase in the period long
119 after trapping (but certainty was low given the wide confidence intervals).

120 The best model for the detection of sugar gliders on cameras also included the interaction between
121 site and time (Table 2). Modeled estimates of the rate of sugar glider detection are presented with

122 95% confidence limits in Figure 2. The low support for the treatment × time period model ($\Delta AICc > 2$;
123 Table 2) allows us to reject the sugar glider lethal control treatment as a potential explanation for
124 the rate of sugar glider detection on cameras. Instead, the model best supported by the data
125 indicates that site level factors were the best predictor of quail egg survival over time. At Southport,
126 sugar gliders were detected more frequently on cameras before and long after the treatment was
127 implemented. There was also a small increase in the rate of sugar glider detections by cameras at
128 Tooms Lake in the long after period. At all other sites, rates of detection were comparable.

129

130 **Discussion**

131 This pilot study shows that site level variation, not the lethal glider control treatment, best explains
132 variation in our quail egg and camera data. Our trapping effort totaled 36 nights, but yielded only
133 nine sugar gliders in total, distributed between the three sites. Given these low capture rates, it is
134 unsurprising that we found no support for the explanatory power of treatment group. Indeed, a real
135 swift parrot nest failed due to sugar glider predation at Southport, indicating that there is a need to
136 rethink the utility of lethal control as a management approach in this study system. The low capture
137 rates we encountered were in spite of intensive trapping effort with limited conservation resources.
138 Whether greater trapping effort, targeted site selection or some other adaptation of our method
139 could improve the efficacy of lethal control remains uncertain, but seems unlikely based on our
140 results, especially in an open system like our study sites.

141 There remain fundamental gaps in knowledge of the ecology of Tasmanian sugar gliders that hinder
142 effective planning for lethal control. Although it is known that the species is widespread in disturbed
143 forests (Allen et al., 2018), there is no information on Tasmanian sugar glider home range size, group
144 size, immigration or behavioural plasticity in the habitats where swift parrots breed. These
145 knowledge gaps are crucial for effective management. Although the scope of our results are limited
146 by small scale and sample size, our pilot study is evidence that reduction of predation on bird nests

147 by gliders is unlikely to be achievable with the approach we trialed. Future studies should evaluate
148 how forest configuration (connectedness to other forest patches) might influence local
149 extinction/recolonization dynamics of sugar gliders (assuming higher capture rates are achievable)
150 and the scale at which trapping needs to occur to affect local densities. Furthermore, the risks
151 identified by Doherty and Ritchie (2017), i.e. herbivore and mesopredator release, disrupted
152 predator social systems, predator immigration and ethical concerns, remain unresolved for our study
153 system. For example, the quail egg predation by common starlings we report highlights that
154 mesopredator release is a possibility because sugar gliders prey on common starlings and potentially
155 suppress their occupancy of nest sites (D.S. unpublished data). Unless common starlings are actively
156 managed concurrently with sugar gliders, any benefits of culling the latter species may be nullified
157 by overabundance of the former. Furthermore, at Tooms Lake (where we trapped the largest
158 number of gliders) we recorded higher rates of quail egg predation and detection of sugar gliders on
159 cameras in the period long after trapping. Whether this is attributable to disruption of the sugar
160 glider social system or immigration is unknown, but has important implications for future lethal
161 control efforts. The major gaps in knowledge of this study system pose non-trivial risks to the
162 effective management of the predation risk to swift parrots using the intensive interventionist
163 approaches we trialed. Earlier work suggests that areas with greater local cover of mature, hollow
164 bearing habitat are at relatively lower risk of predation (Stojanovic et al., 2014). In light of our study,
165 protecting this habitat may be more cost effective over the long term than intensive lethal control at
166 small to medium scales.

167 However sometimes targeted suppression of problematic species can be beneficial in open systems
168 (Crates et al., 2018, Crates et al., 2020), and the results of intervention should be interpreted in
169 context of the effort invested in control/ removal. In our case, we argue that the relative costs of
170 direct interventions like lethal control should be weighed against more general interventions like
171 protection of key breeding sites against ongoing deforestation (Webb et al., 2019). Furthermore, we
172 show that culling gliders may yield no benefit to swift parrots if common starlings are present in an

173 area and exert strong competition for nesting sites. Finally but importantly, there are major
174 unresolved issues surrounding the social license of lethal control of sugar gliders. Given our results,
175 further trials of lethal control for sugar gliders must carefully consider the risks to social license if
176 potential benefits are either difficult to demonstrate or non-existent.

177 Given the ethical and welfare implications of lethal control, we argue that better evidence is needed
178 to support the implementation of culling as a management tool for sugar gliders in Tasmanian
179 forests. Culling wildlife is an important management strategy for managing conservation problems in
180 situ, but requires careful evaluation of outcomes to be justifiable (Salo et al., 2010). Our pilot study is
181 evidence of the value of trialing management techniques to evaluate whether they can achieve
182 conservation goals. Our results are in line with those of other studies that suggest that lethal control
183 does not always have the desired impact if the methods are inefficient or affect a too small fraction
184 of the predator population (Salo et al., 2010, Kämmerle et al., 2019, Cobden et al., 2020). To be
185 effective, integration of multiple management actions targeting different aspects of a conservation
186 problem simultaneously may be necessary (Doherty and Ritchie, 2017).

187 **Conflicts of Interest**

188 The authors declare no conflicts of interest

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241

242 **TABLES**

243 **Table 1.** Ranked model list by AICc for the survival of quail eggs.

Model	df	AICc	ΔAICc	Weights
site × time period	15	415.90	0	1
treatment × time period	6	438.23	22.33	0
treatment	2	454.15	38.25	0
site	5	459.64	43.73	0
null	1	459.71	43.80	0

244

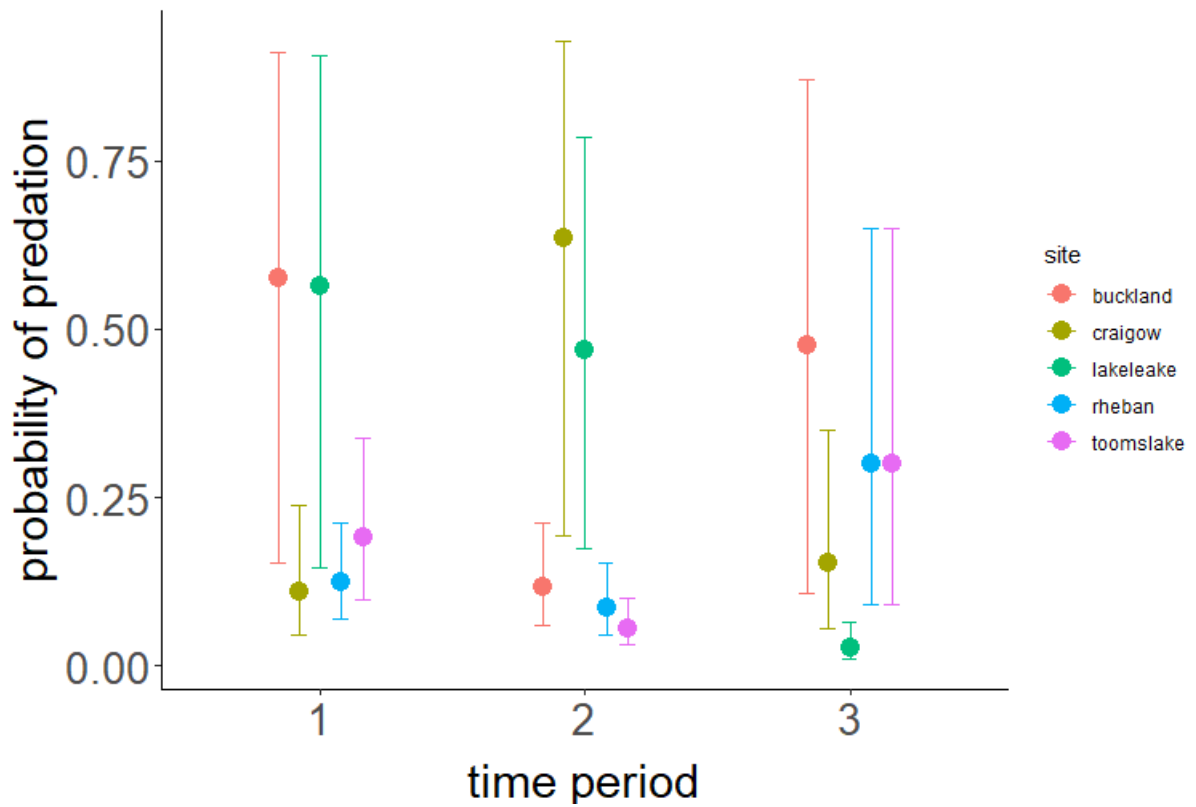
245

246 **Table 2.** Ranked model list by AICc for the frequency of detection of sugar gliders on cameras.

Model	df	AICc	ΔAICc	Weights
site × time period	18	378.12	0	1
treatment × time period	6	461.19	83.07	0
site	5	503.92	125.81	0
treatment	2	572.14	194.03	0
null	1	581.76	203.64	0

247

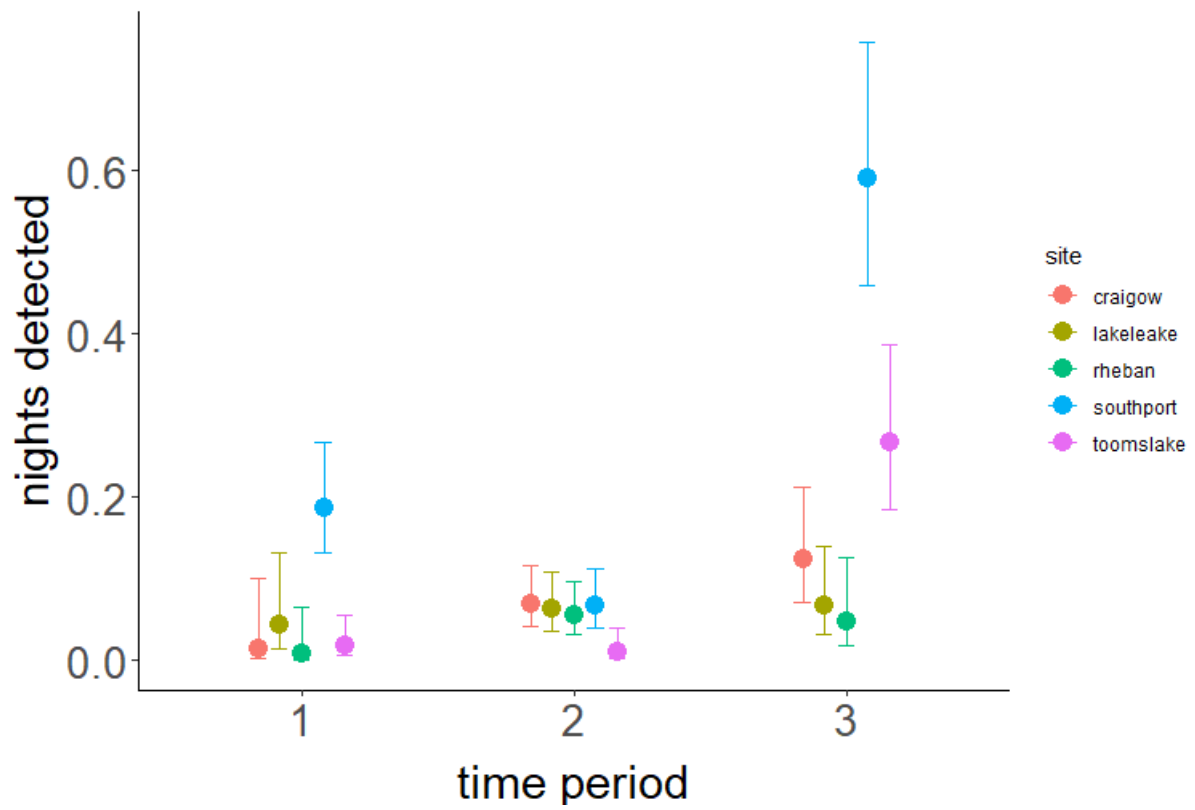
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251 **Figure 1.** Modeled estimates (with 95% confidence limits) of the probability that a quail egg would
252 not survive each time period. Treatment sites were Tooms Lake and Rheban, the others were
253 controls. Time period corresponds to 1 – before, 2 – after, and 3 – long after implementation of the
254 treatment.

255



256

257 **Figure 2.** Modeled estimates (with 95% confidence limits) of the rate at which sugar gliders were
 258 detected on cameras. Treatment sites were Tooms Lake and Rheban and Southport, the others were
 259 controls. Time period corresponds to 1 – before, 2 – after, and 3 – long after implementation of the
 260 treatment.

261