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1 **Low light levels increase avoidance behaviour of diurnal fish species: Implications for**
2 **road culverts.**

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12

13 **Abstract**

14 Inadequately designed culverts are known to pose hydraulic barriers to fish passage,
15 but they may also be behavioural barriers if they adversely affect light levels within them. To
16 test this, we performed a choice experiment and quantified the amount of time individuals of
17 four Australian fish species spent in darkened and illuminated areas of an experimental
18 swimming flume. Behavioural responses were reflective of the species' diel activity patterns;
19 diurnal species preferred illuminated regions, while nocturnal species preferred the darkened
20 region. We then determined a threshold light level of only ~100-200 lux (c.f. midday sunlight
21 ~100,000 lux) was required to overcome the behavioural barrier in ~ 70% of the diurnal fish
22 tested. Placing these threshold values into field context, 100% of culverts sampled recorded
23 inadequate light levels. Attention is required to better understand the impacts of low light
24 levels in culverts on fish passage and to prioritise restoration.

25

26 Keywords: *culvert, fish behaviour, fish passage, fishways, light barrier, small-bodied fish*

27

28 **Introduction**

29 Comprising less than one percent of surface waters, freshwater ecosystems support
30 approximately half of all extant fish species (Reid *et al.* 2013). Yet, competition for, and
31 misuse of, freshwater resources has led to a significant decline in fish diversity and
32 abundance with approximately one third of assessed freshwater fish now at risk of extinction
33 (Dudgeon *et al.* 2006; IUCN 2019). A loss of connectivity between freshwater environments
34 (fragmentation) has been a significant contributor to freshwater fish declines (Baumgartner *et*
35 *al.* 2014; Harris *et al.* 2017; Grill *et al.* 2019). Connectivity is intrinsically linked with access
36 to resources (food, habitat), key population drivers (immigration, emigration, access to
37 spawning grounds), predator avoidance (Harris *et al.* 2017; Rodgers *et al.* 2014; Watson *et al.*
38 2018), and increasingly with climate change, to find refuge pools during drought and to
39 recolonise suitable habitat once flows return. A leading cause of freshwater habitat
40 fragmentation is waterway infrastructure such as dams, weirs and road culverts (Grill *et al.*
41 2019). Traditionally, culverts were designed to move water underneath civil structures in an
42 efficient and cost-effective manner, with little consideration given to the movement
43 requirements of instream biota. Culverts can pose a physical barrier to fish movement by
44 generating excessively high water velocities, excessive turbulence, and by creating a physical
45 jump/drop through bed scouring (Goodrich *et al.* 2018; Rodgers *et al.* 2014; Watson *et al.*
46 2018). Additionally, culverts can act as behavioural barriers if conditions in and around the
47 structure act to dissuade fish from passing through.

48 An emerging concern for fish passage is the potential for altered light levels (i.e. low
49 light during the day or artificial light at night) in and around man-made structures to
50 negatively influence fish movement and behaviour (Jones *et al.* 2017; Perkin *et al.* 2011). For
51 most fish, vision is an important aspect of their sensory repertoire; visual systems are
52 essential for orientation, breeding, foraging and predator avoidance. Fish behaviour is linked
53 with diel light cycles, and it is increasingly apparent that anthropogenic disturbances to
54 natural lighting regimes can have detrimental impacts on affected fish populations (Becker *et al.*
55 *al.* 2013). Several studies have shown that artificial lighting at night (i.e. from street lights,
56 transport networks, industry) can influence reproduction, community structure and movement
57 in nocturnal fish (Becker *et al.* 2013; Riley *et al.* 2012; Ryer *et al.* 2009). Likewise, structures
58 that limit natural light penetration can also alter the behaviour of diurnally active fish (Jones
59 *et al.* 2017).

60 Although light is an important cue regulating the movement behaviour of many fish
61 species, especially salmonids, the specific effects of light on fish passage are highly species-,
62 life stage- and site-specific, and are often influenced by the presence of other behavioural or
63 hydrodynamic stimuli (Banks 1969; Mueller and Simmons 2008; Vowles and Kemp 2012).
64 Several reports have found that low light levels in covered structures (e.g. culverts, fishways,
65 and weirs) can contribute to increased avoidance behaviour during the daytime downstream
66 movements of salmon smolt (Kemp *et al.* 2006; Kemp *et al.* 2005; Kemp and Williams 2008;
67 Tétard *et al.* 2019; Welton *et al.* 2002). Similarly, avoidance of darkened environments
68 within covered fishways contributes to the reduced movement of several small-bodied
69 Australian freshwater fish species (Jones *et al.* 2017). Abrupt changes in light intensity such
70 as at the entrance/exits of fishways or culverts can also cause avoidance behaviour in
71 lampreys (Moser and Mesa 2009) and juvenile salmon (Ono and Simenstad 2014). However,
72 other studies have demonstrated that upstream migrating salmon, trout, eels, Topeka Shiner,

73 Fathead Minnow and common galaxias in Australia, are unaffected by reduced light levels in
74 civil structures (Fjeldstad *et al.* 2018; Amtstaetter *et al.* 2017; Kozarek *et al.* 2017; Gowans *et*
75 *al.* 2003; Rogers and Cane 1979). These conflicting accounts of the effect of light on fish
76 movement suggest a range of species-specific behavioural responses to different lighting
77 conditions (Fjeldstad *et al.* 2018; Amtstaetter *et al.* 2017; Kozarek *et al.* 2017; Gowans *et al.*
78 2003; Rogers and Cane 1979), with such variability also indicating that our understanding of
79 the effects of altered lighting regimes on fish movement is poor, despite this issue being
80 raised in several fish passage guidelines (e.g. Fairfull and Witheridge 2003; Franklin *et al.*
81 2018).

82 Accordingly, research is required to better understand the potential for low light levels
83 within culverts to impact fish movement and behaviour to inform the regulation of new
84 culvert structures and to guide the remediation of existing structures. The aim of this study
85 was to quantify the effect of reduced light levels on the movement behaviour of four species
86 of small-bodied or juvenile Australian native fish. We chose two small-bodied species, Fly-
87 specked Hardyhead (*Craterocephalus stercusmuscarum*) (Günther, 1867) and Australian
88 Smelt (*Retropinna semoni*) (Weber, 1895), both of which have maximum adult sizes of 7 cm.
89 We also included juveniles of two large-bodied species, Australian Bass (*Macquaria*
90 *novemaculeata*) (Steindachner 1866) and Silver Perch (*Bidyanus bidyanus*) (Mitchell 1838)
91 that have respective maximum adult sizes of 60 and 40 cm. Three of the species, Fly-specked
92 Hardyhead, Australian Smelt and Silver Perch, are more active during the daytime
93 (Baumgartner *et al.* 2008; Clunie and Koehn 2001; Mallen-Cooper 1999; Stuart and Mallen-
94 Cooper 1999), while Australian Bass are generally crepuscular but can be active at other
95 times of the day and night (Harris 1985; Smith *et al.* 2011). We hypothesised that Fly-
96 specked Hardyhead, Australian Smelt and Silver Perch would prefer an illuminated
97 environment, and that Australian Bass would prefer a darker environment. We then aimed to

98 establish the minimum lighting thresholds for the species that displayed a preference for an
99 illuminated environment. Finally, we placed these light threshold values into context by
100 comparing them with light levels measured within existing culverts in south-east Queensland,
101 Australia.

102

103 **Methods**

104 *Fish collection and husbandry*

105 Juvenile Australian Bass ($n = 40$; TL: mean \pm SD 73.4 ± 8.7 mm;) and Silver Perch (n
106 $= 70$; mean \pm SD 65.95 ± 16.6 mm) were sourced from commercial hatcheries. Adult Fly-
107 specked Hardyheads ($n = 110$; mean \pm SD 48.9 ± 4.7 mm) were supplied by a commercial
108 collector (Aquadreen, Howard Springs, Northern Territory), from the Howard River,
109 Girraween Road Crossing, Northern Territory ($12^{\circ}31'51''S$ $131^{\circ}07'41''E$). Adult Smelt ($n =$
110 60 ; mean \pm SD 42.45 ± 6.5 mm) were collected using nets at Cedar Creek and Moggil Creek,
111 Brisbane, Queensland ($27^{\circ}19'28.6''S$ $152^{\circ}47'39.1''E$ and $27^{\circ}30'16.1''S$ $152^{\circ}55'50.1''E$,
112 respectively).

113 The fish were housed at the Biohydrodynamics Laboratory at the University of
114 Queensland (Brisbane, Queensland, Australia). Fish were kept with conspecifics in 40 L glass
115 aquaria that formed part of a 1000 L recirculating system with mechanical and biological
116 filtration and UV sterilization. The water temperature was maintained at $25^{\circ}C \pm 1^{\circ}C$. Fish
117 were fed commercial aquaculture pellets (Ridley, Brisbane Australia) and exposed to a 12-
118 hour light-dark cycle provided by overhead LED aquarium lighting. The ambient light
119 intensity was measured at the water level of the housing aquaria using a photometer (Extech
120 HD450, New Hampshire, U.S.A.), which averaged 2535 ± 238.6 lux (mean \pm SD).

121

122 *Light – dark behavioural trials*

123 Behavioural trials were performed on all four species in a 12-metre hydraulic channel
124 (12.0 x 0.5 x 0.3 m). The light around and within the channel was controlled using blackout
125 plastic sheeting to create an environment with zero ambient light (0 lux). The integrity of this
126 screen was checked before starting trials each day to ensure no external light sources were
127 present. Half of the channel was illuminated using 4000 K correlated colour temperature LED
128 lighting (Atom 56-watt batten, China) and the other half left darkened. The light intensity
129 above the illuminated half was set to 2535 lux, the same as above the housing aquaria. A
130 sharp light-dark transition point was achieved by dividing the darkened area around the
131 channel with black plastic. This included the space within the channel above the waterline.

132 Four treatments were required to control for the direction of water flow in the channel
133 that could not be changed, and the illuminated state (light or dark) of the release point (Fig.
134 1). The first two treatments were with the downstream half of the channel illuminated and the
135 upstream dark. Ten fish per species were randomly allocated to each treatment for each trial.
136 The time each fish spent in each zone of the channel was observed for 30 min through a small
137 observation point at the transition zone between the illuminated and darkened areas. All fish
138 were swum individually and were released 1 m from either end of the channel, facing into the
139 water flow (Fig. 1). The channel bulk velocity was set to 0.3 m s⁻¹ and the depth set to 0.15
140 m, measured at the mid-point, 6 m along the channel length. This velocity was chosen as it
141 was significantly below the maximum sustainable swimming speed (U_{crit} ; Brett 1964) of all
142 four species tested (Watson *et al.* 2019) so as to minimise any effect that swimming capacity,
143 or their innate rheotactic response, could have on their subsequent behaviour. The water
144 temperature was maintained at $25 \pm 1^\circ\text{C}$.

145

146 *Determining the light intensity thresholds for fly-specked hardyheads and smelt*

147 Of the four species tested, the Fly-specked Hardyhead and Australian Smelt displayed
148 strong avoidance of darkened environments which could negatively impact their movement
149 through artificially darkened culverts. To understand the minimum illumination levels that
150 would encourage Fly-specked Hardyhead and Australian Smelt to enter a darkened
151 environment, we set up the flume with the downstream half illuminated and upstream half
152 darkened, and sequentially increased the light levels in the darkened section of the flume. The
153 illuminated zone was set to the same light intensity as the housing aquaria (2535 lux). The
154 darkened half of the channel was fitted with an overhead controllable LED strip light (ML-
155 1009FAWi, MELEC, Birtinya, Queensland, Australia) that allowed us to incrementally
156 increase light intensities in the darkened region. The light intensity treatments were
157 determined by the response times as the experiment progressed. Fly-specked Hardyhead were
158 released at 5, 10, 25, 50, 250, 300 and 400 lux, and Australian Smelt were released at 2.5, 5,
159 25 and 200 lux. We recorded the total time fish spent in both halves of the channel over 10
160 min, and the proportion (as a percentage) of individuals that used the darkened region of the
161 flume (for any length of time). Fifteen fish of each species were individually tested in each
162 light intensity treatment and released 1 m from the downstream end of the flume. Individual
163 fish were only swum once. This was not done for Australian Bass which preferred the dark, or
164 Silver Perch that showed no light-dark preference.

165

166 *Sampling light levels of culverts*

167 To place the light intensity threshold values obtained for Fly-specked Hardyhead and
168 Australian Smelt into context, we sampled the ambient light levels within and outside fifteen
169 culverts within south-east Queensland (Australia) using a photometer (Extech HD450, New

170 Hampshire, U.S.A.). Sampling was undertaken between 09:00 and 12:15 on the 16 December
171 2018 (austral summer), on a cloudless day when ambient light levels within the culvert would
172 be at, or close to, their maximum levels. The culverts sampled were predominantly dual
173 carriage roadways and one single pedestrian crossing (culvert range 3.4 – 7.0 m in length, ~
174 1.0 m height). All culverts contained at least 0.2 m water depth at the time of sampling.

175

176 *Data analyses*

177 All statistical analyses were performed using R version 1.1.423 (R Core Team 2017)
178 in the RStudio environment. The preference experiment data was analysed using a
179 quasibinomial generalised linear model and ANOVA with species, release condition (light or
180 dark) and release point (downstream or upstream) as predictors, allowing for possible
181 interactions. To analyse how many fish were entering the treatment zone, a binomial linear
182 regression was fit to the data with ‘entering’ (y/n) as the response variable, and the time of
183 day the fish were swum, species, light level in the treatment zone (lux) and fish length as
184 predictors allowing all interactions. A subsequent ANOVA revealed that the light level in the
185 treatment zone was the only significant predictor and the model was reduced accordingly.
186 Statistical significance for all analysis was set at $P < 0.05$.

187

188 **Results**

189 *Light – dark preferences*

190 There was a statistically significant 3-way interaction between species, release point
191 and lighting condition of the release point ($F_{3, 144} = 11.2284, p < 0.001$) due to the behaviour
192 of Silver Perch (Fig. 2). When released downstream in the darkened environment, Silver
193 Perch swam upstream into the illuminated zone, and when released downstream in the light

194 they spent they spent around half of their time in the light. When released upstream they were
195 indifferent in their lighting preference and stayed in the upstream zone.

196 Australian Bass appeared to prefer the darkened zone of the channel, spending on
197 average 91.3% of their time there across all treatments. Irrespective of flow direction,
198 individual Australian Bass that were released in the illuminated zone (treatments 2 and 3)
199 rapidly moved to the darkened zone. Neither the release point, nor the illumination condition
200 at the release point were found to affect the time spent in either the light or dark zones.

201 In contrast, the Fly-specked Hardyhead and Australian Smelt both displayed a strong
202 avoidance of the darkened zone (or preference for the light). Fly-specked Hardyhead and
203 Australian Smelt were observed spending respectively 97.2% and 86.3% of their total trial
204 time across all treatments in the illuminated zone. Like Australian Bass, neither the release
205 point, nor the illumination condition at the release point were found to affect the time spent in
206 either the light or dark zones. We observed that both the Fly-specked Hardyhead and
207 Australian Smelt quickly moved to the illuminated zone of the channel when released in the
208 darkened zone.

209

210 *Light intensity thresholds stimulating fish movement*

211 Given that both Fly-specked Hardyhead and Australian Smelt displayed strong
212 avoidance of the darkened environment in the channel, we gradually increased the
213 illumination in the darkened (treatment) zone to determine the light threshold that would
214 encourage these species to enter. Overall, the number of individuals entering the treatment
215 region of the channel increased with increasing illumination ($F_{(1, 168)} = 28.921, p < 0.001$)
216 (Fig. 3). It is worth noting that while fish length did not have a statistically significant effect

217 on the number of individuals entering the darkened treatment zones, length is potentially
218 biologically significant with more larger fish entering at lower light levels ($p = 0.056$).

219 Fly-specked Hardyhead began entering the darkened region of the flume at 5 lux, with
220 26% of individuals observed entering. The number of individuals entering the darkened
221 region of the flume remained at less than 50% until illumination levels exceeded 200 lux. The
222 illumination threshold at which smelt started to enter the darkened region was 2.5 lux (13%
223 of individuals). Doubling the amount of available light from 2.5 to 5 lux resulted in a four-
224 fold increase to 53% of individuals entering. Further increasing the light intensity to 25 lux
225 resulted in more than 75% of individuals entering the treatment zone.

226

227 *Light intensity responses in field context*

228 We quantified the illumination levels in 15 culverts in south-east Queensland to
229 determine how many reached the minimum lighting thresholds required to encourage 70% of
230 Australian Smelt and Fly-specked Hardyhead to successfully move into a darkened
231 environment. The modelled threshold values corresponded to 100 and 200 lux for Australian
232 Smelt and Fly-specked Hardyhead, respectively. We found that lighting levels at the culvert
233 entrance/exit averaged 70.9 ± 44.8 lux (mean \pm s.d.; range: 5.6 - 123.1 lux; Table 1). In all
234 culverts, light levels dropped to less than 3 lux in the centre (0.6 ± 0.8 lux; range 0 – 2.3 lux).
235 Based on the light threshold determined for both Australian Smelt and Fly-specked
236 Hardyhead, all culverts sampled contained insufficient light in the centre to promote a 70%
237 passage success rate.

238

239 **Discussion**

240 Here we show that the levels of available light significantly affected the behaviour of
241 three out of the four Australian fish species examined, and that our results were mostly
242 consistent with the hypothesis that species behavioural responses to lighting levels would
243 relate to their daily activity patterns (i.e. diurnal versus nocturnal). The largely diurnal
244 Australian Smelt and Fly-specked Hardyhead showed near absolute avoidance of the
245 completely darkened environment within the experimental channel, while the nocturnal
246 Australian Bass strongly avoided the illuminated section. Surprisingly, Silver Perch showed
247 no preference for either the illuminated or darkened environment. Silver Perch activity
248 patterns in the wild are generally greatest during daylight hours, however, they do not appear
249 to be actively inhibited by darkness and can be trapped, albeit at lower frequencies, at night
250 (Baumgartner *et al.* 2008). In the present study, we were unable to disentangle a behavioural
251 response to light levels from their response to water flow direction (rheotaxis), which may
252 have been exacerbated by the relatively slow flow velocities used in this study compared to
253 their swimming capacity (Watson *et al.* 2019). For the two species that avoided the darkened
254 environment, we found that the threshold light intensities needed to encourage individual fish
255 to enter the darker half of the test channel were quite low. These data also suggest that
256 providing even very low levels of light with artificial lighting could remove the behavioural
257 barrier that culverts may pose to diurnally active fish species.

258 The behavioural response of Australian Smelt, Fly-specked Hardyhead and Australian
259 Bass provided an indication of the range of responses of fish species to low light levels
260 representative of those within culverts, and how broad diel classifications can help to predict
261 behavioural responses of those species most at risk of low light levels. Yet, consideration
262 must be given to factors other than diel classification, such as movement motivation.
263 Diadromous species that are obligate migrators for example, may be more likely to pass
264 through a darkened culvert due a fundamental requirement to reach the sea or freshwater, as

265 compared with facultative migrators. Indeed, the movement behaviour of *Galaxias spp.* was
266 unaffected by a 70 m long darkened (0 lux) pipe culvert along an upstream migration path
267 (Amtstaetter *et al.* 2017). In contrast, facultative migrators such as smelt, may lack that
268 motivation and are more susceptible to altered light regimes (Jones *et al.* 2017). Clearly fish
269 species differ in their readiness to use low light environments and so appropriate
270 consideration of interspecific differences and variation in their tolerance of darkness within
271 man-made structures needs to be given.

272 Abrupt lighting changes at sharp transitional point (as opposed to a graded transition)
273 can be why some fish display strong behavioural reactions to light levels in some fish passage
274 structures. Some fish may avoid areas where shadows cast by anthropogenic structures cause
275 abrupt lighting changes because of the risk of predators using the shaded areas as cover
276 (Kemp *et al.* 2005; Ono and Simenstad 2014; Steenbergen *et al.* 2011). However, data from
277 our study, which also employed a sharp transition from light to dark, suggests this may not
278 have been a constraining factor influencing the movement of the four fish species we
279 examined. We found that amongst the species that showed a distinct light-dark preference
280 response, nearly all individuals rapidly moved to their preferred illumination zone, regardless
281 of the flow orientation or illumination state at the release point. While the sharp light gradient
282 did not appear to completely restrict their initial movement into or out of the dark zone,
283 further work will be required to determine if the abrupt light-dark transition influenced
284 subsequent use of the space by the fish.

285 To determine the prevalence of prohibitively low light levels for fish passage in
286 culverts, we measured light levels in 15 box or pipe culverts in Brisbane, Australia ranging
287 from 3-8 m in length. We found that light levels at both the entrance and exit of the culvert
288 ranged from ~5 to ~120 lux. Less than 3 lux was recorded in the middle of all culverts
289 irrespective of culvert length. Based on the lighting thresholds for Fly-specked Hardyhead

290 and Australian Smelt, all of the culverts examined could pose as a behavioural barrier to these
291 species. Although we only measured light levels on just one day at each culvert, we measured
292 at the brightest time of day (morning) and year (summer), so if prohibitively low light levels
293 were detected under these conditions, they are likely to be light barriers at other times of year.
294 The amount of ambient light present within the structure is dependent upon a culverts' cross-
295 sectional area, height and orientation relative to a light source. Environmental factors such as
296 season, water depth and surrounding riparian vegetation density, will also affect the amount
297 of light within a culvert, which means that the level of ambient light in a structure may vary
298 considerably over daily and seasonal scales. A greater understanding of how culvert lighting
299 conditions change over time is important for determining if and when a particular structure is
300 likely to be a behavioural movement barrier for fish. Culvert lighting requirements need to be
301 considered in the context of other culvert design features to ensure that efforts to mitigate
302 hydrological barriers to fish passage (e.g. significant slope or excessive water velocities) do
303 not inadvertently create behavioural obstacles to fish passage.

304 When assessing if lighting levels are likely to influence fish passage through culverts,
305 it is important to consider other factors that may influence fish behaviour such as the
306 presence of predators and food (Magurran 1990; Morgan and Godin 1985), schooling effects
307 with conspecifics (Krause *et al.* 2000) or individuals' personality (Hirsch *et al.* 2017). Our
308 study was conducted under controlled laboratory conditions focusing on the test species'
309 response to light levels. Progressively overlaying the levels of complexity found in the field
310 would strengthen our understanding of when reduced light levels are barriers, and how they
311 may be overcome. For example, increases in water flow may reach a threshold point where a
312 low light level ceases to be a behavioural barrier simply due to a strengthening of the fishes
313 rheotactic response. Finally, light pollution at night from anthropogenic sources (Holker *et al.*

314 2010; Perkin *et al.* 2014) should be considered as increased levels of artificial ambient light
315 along waterways may influence the movement behaviour of nocturnal species.

316 Currently, many fish passage guidelines for road crossing structures identify low light
317 as a potential barrier for fish passage and recommend that light levels be considered by
318 infrastructure planners and asset owners (Fairfull and Witheridge 2003; Franklin *et al.* 2018).
319 Where light levels in culverts are predicted to impede passage of target fish species, alternate
320 road crossing structures (such as bridges) may be recommended. However, in circumstances
321 where low daytime light levels within a culvert may be unavoidable, our data suggest that
322 providing small amounts of light through the installation of artificial lights or the provision of
323 skylights, could remove a behavioural obstacle for some diurnally active fish species. More
324 data on lighting thresholds for the movement of a greater range of fish species will inform
325 fish passage guidelines and allow recommendations to be made in a site- and species-specific
326 manner dictated by the culvert length, orientation, and the passage requirements of the local
327 fish community.

328

329 **Conclusions**

330 Our study showed that light levels affected the movement behaviour of three
331 Australian native fish species, and that low light levels impeded the movement of two of the
332 four species in an experimental channel. Optimal lighting levels for fish passage should be
333 considered in the future design of artificial instream structures such as culverts and fishways,
334 and in the remediation of existing structures. Our results indicate that only relatively low
335 ambient daytime light levels are required within closed structures to encourage movement by
336 certain diurnal species, and that fish willingly move into a darkened environment with the
337 provision of artificial light. Developing minimum lighting standards that take into account

338 species-specific light requirements can lead to improved passage rates through culverts,
339 reduced fragmentation, and more resilient fish populations.

340

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347

348 **References**

349 Amtstaetter, F., O'Connor, J., Borg, D., Stuart, I., and Moloney, P. (2017) Remediation of
350 upstream passage for migrating *Galaxias* (Family: Galaxiidae) through a pipe culvert.
351 *Fisheries Management and Ecology* **24**(3), 186-192.

352

353 Banks, J.W. (1969) A Review of the literature on the upstream migration of adult salmonids.
354 *Journal of Fish Biology* **1**(2), 85-136.

355

356 Baumgartner, L., Zampatti, B., Jones, M., Stuart, I., and Mallen-Cooper, M. (2014) Fish
357 passage in the Murray-Darling Basin, Australia: Not just an upstream battle. *Ecological*
358 *Management & Restoration* **15**(1), 28-39.

359

360 Baumgartner, L.J., Stuart, I.G., and Zampatti, B.P. (2008) Determining diel variation in fish
361 assemblages downstream of three weirs in a regulated lowland river. *Journal of Fish Biology*
362 **72**(1), 218-232.

363

364 Becker, A., Whitfield, A.K., Cowley, P.D., Jarnegren, J., and Naesje, T.F. (2013) Potential
365 effects of artificial light associated with anthropogenic infrastructure on the abundance and
366 foraging behaviour of estuary-associated fishes. *Journal of Applied Ecology* **50**(1), 43-50.

367

368 Brett, J.R. (1964) The respiratory metabolism and swimming performance of young Sockeye
369 salmon. *Journal of the Fisheries Research Board of Canada* **21**(5), 1183-1226.

370

371 Clunie, P., and Koehn, J.D. (2001) silver perch: A Resource Document. Arthur Rylah
372 Institute for Environmental Research Department of Natural Resources and Environment,
373 Melbourne.

374

375 Dudgeon, D., Arthington, A.H., Gessner, M.O., Kawabata, Z.-I., Knowler, D.J., Lévêque, C.,
376 Naiman, R.J., Prieur-Richard, A.-H., Soto, D., Stiassny, M.L.J., and Sullivan, C.A. (2006)
377 Freshwater biodiversity: importance, threats, status and conservation challenges. *Biological*
378 *Reviews* **81**(02), 163.

379

380 Fairfull, S., and Witheridge, G. (2003) Why do fish need to cross the road? Fish passage
381 requirements for waterway crossings. NSW Fisheries, Cronulla, 16 pp.

382

383 Fjeldstad, H.-P., Pulg, U., and Forseth, T. (2018) Safe two-way migration for salmonids and
384 eel past hydropower structures in Europe: a review and recommendations for best-practice
385 solutions. *Marine and Freshwater Research* **69**(12), 1834-1847.

386

387 Franklin, P., Gee, E., Baker, C., and Bowie, S. (2018) New Zealand Fish Passage Guidelines
388 for Structures up to 4m. National Institute of Water & Atmospheric Research Ltd, Hamilton,
389 NZ.

390

391 Goodrich, H.R., Watson, J.R., Cramp, R.L., Gordos, M.A., and Franklin, C.E. (2018) Making
392 culverts great again. Efficacy of a common culvert remediation strategy across sympatric fish
393 species. *Ecological Engineering* **116**, 143-153.

394

395 Gowans, A.R.D., Armstrong, J.D., Priede, I.G., and Mckelvey, S. (2003) Movements of
396 Atlantic salmon migrating upstream through a fish-pass complex in Scotland. *Ecology of*
397 *Freshwater Fish* **12**(3), 177-189.

398

399 Grill, G., Lehner, B., Thieme, M. et al. Mapping the world's free-flowing rivers. *Nature* **569**,
400 215–221 (2019) doi:10.1038/s41586-019-1111-9

401

402 Günther, A. (1867) Additions to the knowledge of Australian reptiles and fishes. *Annals and*
403 *Magazine of Natural History* (3)20(8): 45-68.

404

405 Harris, J.H. (1985) Diet of the Australian bass, *Macquaria novemaculeata* (Perciformes,
406 Percichthyidae), in the Sydney Basin. *Australian Journal of Marine and Freshwater*
407 *Research* **36**(2), 219-234.

408

409 Harris, J.H., Kingsford, R.T., Peirson, W., and Baumgartner, L.J. (2017) Mitigating the
410 effects of barriers to freshwater fish migrations: the Australian experience. *Marine and*
411 *Freshwater Research* **68**(4), 614-628.

412

413 Hirsch, P.E., Thorlacius, M., Brodin, T., and Burkhardt-Holm, P. (2017) An approach to
414 incorporate individual personality in modeling fish dispersal across in-stream barriers.
415 *Ecology and Evolution* **7**(2), 720-732.

416

417 Holker, F., Wolter, C., Perkin, E.K., and Tockner, K. (2010) Light pollution as a biodiversity
418 threat. *Trends in Ecology & Evolution* **25**(12), 681-682.

419

420 IUCN (2019) The IUCN Red List of Threatened Species. Version 2019-1.
421 <http://www.iucnredlist.org>. Downloaded on 21 March 2019.

422

423 Jones, M.J., Baumgartner, L.J., Zampatti, B.P., and Beyer, K. (2017) Low light inhibits
424 native fish movement through a vertical-slot fishway: Implications for engineering design.
425 *Fisheries management and ecology* **24**(3), 177-185.

426

427 Kemp, P.S., Gessel, M.H., Sandford, B.P., and Williams, J.G. (2006) The behaviour of
428 Pacific salmonid smolts during passage over two experimental weirs under light and dark
429 conditions. *River Research and Applications* **22**(4), 429-440.

430

431 Kemp, P.S., Gessel, M.H., and Williams, J.G. (2005) Seaward migrating subyearling chinook
432 salmon avoid overhead cover. *Journal of Fish Biology* **67**(5), 1381-1391.

433

434 Kemp, P.S., and Williams, J.G. (2008) Response of migrating Chinook salmon
435 (*Oncorhynchus tshawytscha*) smolts to in-stream structure associated with culverts. *River*
436 *Research and Applications* **24**(5), 571-579.

437

438 Kozarek, J., Hatch, J., and Mosey, B. (2017) Culvert length and interior lighting impacts to
439 Topeka Shiner passage. Final Report 2014-2017. Minnesota Department of Transportation,
440 St. Paul, Minnesota

441

442 Krause, J., Butlin, R.K., Peuhkuri, N., and Pritchard, V.L. (2000) The social organization of
443 fish shoals: a test of the predictive power of laboratory experiments for the field. *Biological*
444 *Reviews of the Cambridge Philosophical Society* **75**(4), 477-501.

445

446 Magurran, A.E. (1990) The inheritance and development of minnow antipredator behavior.
447 *Animal Behaviour* **39**, 834-842.

448

449 Mallen-Cooper, M. (1999) Developing fishways for nonsalmonid fishes: a case study from
450 the Murray River in Australia. In *Innovations in Fish Passage Technology*. (Ed. M Odeh) pp.
451 173-195. (American Fisheries Society: Bethesda, Maryland)

452

453 Mitchell, T. (1838) *Three Expeditions into the Interior of Eastern Australia; with*
454 *Descriptions of the Recently Explored Region of Australia Felix, and of the Present Colony*
455 *of New South Wales* (2nd ed.). London: T. and W. Boone.

456

457 Morgan, M.J., and Godin, J.G.J. (1985) Antipredator benefits of schooling behavior in a
458 cyprinodontid fish, the banded killifish (*Fundulus diaphanus*). *Journal of Comparative*
459 *Ethology* **70**(3), 236-246.

460

461 Moser, M.L., and Mesa, M.G. Passage considerations for anadromous lampreys. In 'Biology,
462 Management, and Conservation of Lampreys in North America. American Fisheries Society
463 Symposium', 2009, pp. 115–124

464

465 Mueller, R.P., and Simmons, M.A. (2008) Characterization of gatewell orifice lighting at the
466 Bonneville Dam second powerhouse and compendium of research on light guidance with
467 juvenile salmonids. Final Report prepared for the U.S. Army Corps of Engineers, Portland
468 District, under a Government Order with the U.S. Department of Energy Contract DE-AC05-
469 76RL01830.

470

471 Ono, K., and Simenstad, C.A. (2014) Reducing the effect of overwater structures on
472 migrating juvenile salmon: An experiment with light. *Ecological Engineering* **71**, 180-189.

473

474 Perkin, E.K., Hölker, F., Richardson, J.S., Sadler, J.P., Wolter, C., and Tockner, K. (2011)

475 The influence of artificial light on stream and riparian ecosystems: questions, challenges, and
476 perspectives. *Ecosphere* **2**(11), art122.

477

478 Perkin, E.K., Hölker, F., Tockner, K., and Richardson, J.S. (2014) Artificial light as a
479 disturbance to light-naïve streams. *Freshwater Biology* **59**(11), 2235-2244.

480

481 R Core Team (2017) R: A language and environment for statistical computing. *R Foundation*
482 *for Statistical Computing*. Vienna, Austria. URL <http://www.R-project.org/>.

483

484 Reid, G.M., Contreras Macbeath, T., and Csatádi, K. (2013) Global challenges in freshwater-
485 fish conservation related to public aquariums and the aquarium industry. *International Zoo*
486 *Yearbook* **47**(1), 6-45.

487

488 Riley, W.D., Bendall, B., Ives, M.J., Edmonds, N.J., and Maxwell, D.L. (2012) Street
489 lighting disrupts the diel migratory pattern of wild Atlantic salmon, *Salmo salar* L., smolts
490 leaving their natal stream. *Aquaculture* **330-333**, 74-81.

491

492 Rodgers, E.M., Cramp, R.L., Gordos, M., Weier, A., Fairfall, S., Riches, M., and Franklin,
493 C.E. (2014) Facilitating upstream passage of small-bodied fishes: linking the thermal
494 dependence of swimming ability to culvert design. *Marine and Freshwater Research* **65**(8),
495 710-719.

496

497 Rogers, A., and Cane, A. (1979) Upstream passage of adult salmon through an unlit tunnel.
498 *Aquaculture Research* **10**(2), 87-92.

499

500 Ryer, C.H., Stoner, A.W., Iseri, P.J., and Spencer, M.L. (2009) Effects of simulated
501 underwater vehicle lighting on fish behavior. *Marine Ecology Progress Series* **391**, 97-106.

502

503 Smith, J.A., Baumgartner, L.J., Suthers, I.M., and Taylor, M.D. (2011) Distribution and
504 movement of a stocked freshwater fish: implications of a variable habitat volume for stocking
505 programs. *Marine and Freshwater Research* **62**(11), 1342-1353.

506

507 Steenbergen, P.J., Richardson, M.K., and Champagne, D.L. (2011) Patterns of avoidance
508 behaviours in the light/dark preference test in young juvenile zebrafish: A pharmacological
509 study. *Behavioural Brain Research* **222**(1), 15-25.

510

511 Steindachner, F. (1866) Anzeiger der Kaiserlichen Akademie der Wissenschaften,
512 Mathematisch-Naturwissenschaftlichen Classe v. 3 (no. 7); Steindachner Abstract Port
513 Jackson, New South Wales, Australia.

514

515 Stuart, I.G., and Mallen-Cooper, M. (1999) An assessment of the effectiveness of a vertical-
516 slot fishway for non-salmonid fish at a tidal barrier on a large tropical/subtropical river.
517 *Regulated Rivers-Research & Management* **15**(6), 575-590.

518

519 Tétard, S., Maire, A., Lemaire, M., De Oliveira, E., Martin, P., and Courret, D. (2019)
520 Behaviour of Atlantic salmon smolts approaching a bypass under light and dark conditions:
521 Importance of fish development. *Ecological Engineering* **131**, 39-52.

522

523 Vowles, A.S., and Kemp, P.S. (2012) Effects of light on the behaviour of brown trout (*Salmo*
524 *trutta*) encountering accelerating flow: Application to downstream fish passage. *Ecological*
525 *Engineering* **47**, 247-253.

526

527 Watson, J.R., Goodrich, H.R., Cramp, R.L., Gordos, M.A., and Franklin, C.E. (2018)
528 Utilising the boundary layer to help restore the connectivity of fish habitats and populations.
529 *Ecological Engineering* **122**, 286-294.

530

531 Watson, J.R., Goodrich, H.R., Cramp, R.L., Gordos, M.A., Yan, Y., Ward, P.J. and Franklin,
532 C.E. (2019) Swimming performance traits of twenty-one Australian fish species: a fish
533 passage management tool for use in modified freshwater systems. *bioRxiv*,
534 doi:10.1101/861898

535

536 Weber, M. (1895) Fische von Ambon, Java, Thursday Island dem Burnett-Fluss und von der
537 Sud-Küste von Neu-Guinea. *Zoological Studies* 5(2):257-276 1 fig.

538

539 Welton, J.S., Beaumont, W.R.C., and Clarke, R.T. (2002) The efficacy of air, sound and
540 acoustic bubble screens in deflecting Atlantic salmon, *Salmo salar* L., smolts in the River
541 Frome, UK. *Fisheries Management and Ecology* **9**(1), 11-18.

543 **Tables**

544 **Table 1:** Light readings recorded at 15 roadway culverts.

Sampling time (hh:mm)	Culvert type	External light level at the culvert entrance (Lux)	Light level in the middle of the culvert (Lux)	External light level at the culvert exit (Lux)	Ambient light outside culvert (Lux)
09:00	Box	114.9	0.2	106.2	72 100
09:20	Box	113.8	0.3	119.7	72 500
09:25	Pipe	92.3	1.8	105.5	73 300
09:30	Pipe	5.6	0	9.4	73 500
09:40	Pipe	11.3	0.1	24.7	73 800
10:05	Box	30.5	0.3	60.1	74 100
10:10	Box	107.1	0.1	112.7	78 300
10:30	Box	114.7	1.2	75.6	81 800
10:35	Box	105.3	0	59.1	84 200
10:40	Pipe	14.5	0	89.2	86 200
11:15	Box	116.7	2.1	106.8	88 100
11:25	Box	121.3	2.3	123.1	89 700
11:35	Box	21.2	0.4	18.5	88 200
12:00	Box	16.1	0.1	18.6	89 500
12:15	Box	12	0.3	101.9	88 200

545

546

547 **Figure Legends**

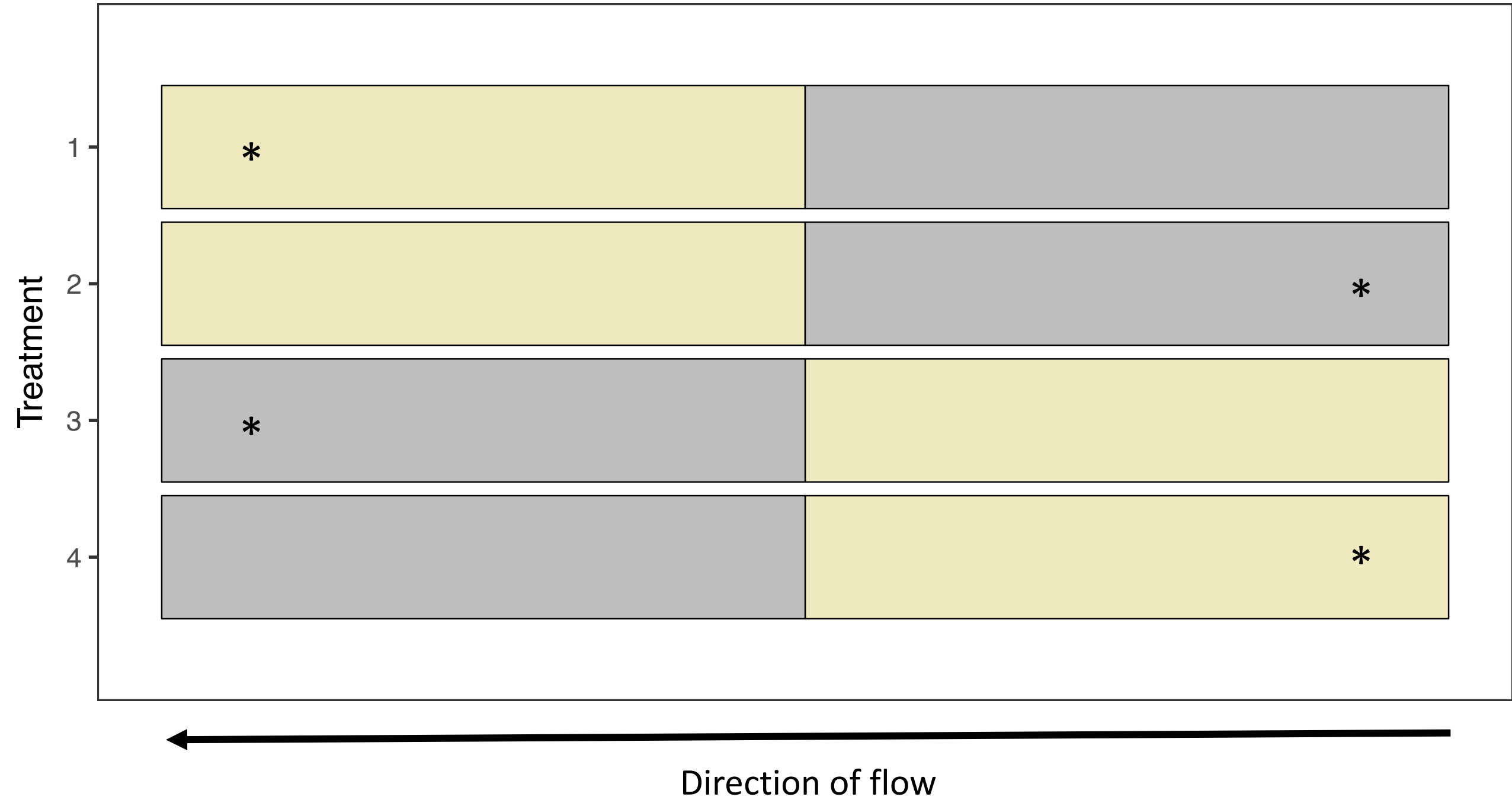
548 **Figure 1.** Schematic of experimental design to assess light preferences. Side view of the 12
549 m long experimental hydraulic flume, indicating the position of release (asterisks) and
550 location of the darkened (grey) and illuminated (yellow) zones in relation to the direction of
551 water flow.

552

553 **Figure 2.** The average time spent in the light half of the experimental channel for each
554 species and treatment, showing the statistically significant 3-way interaction was due to the
555 behaviour of the Silver perch. Fly-specked hardyheads and smelt showed a strong preference
556 for the illuminated region of the channel regardless of release point or its lighting conditions.
557 Conversely Australian bass strongly preferred a dark environment regardless of release point
558 or its lighting condition. Error bars represent 95% confidence intervals.

559

560 **Figure 3.** Regression curves showing the probability of a fish entering the darkened half of the
561 experimental channel with increasing light levels. Both species were modelled individually and
562 combined. Smelt were only tested up to 200 lux. Error bars represent 95% confidence intervals.



Release point

Downstream

Upstream

Proportion of time spent in light

1.00
0.75
0.50
0.25
0.00

Dark

Light

Dark

Light

Release condition

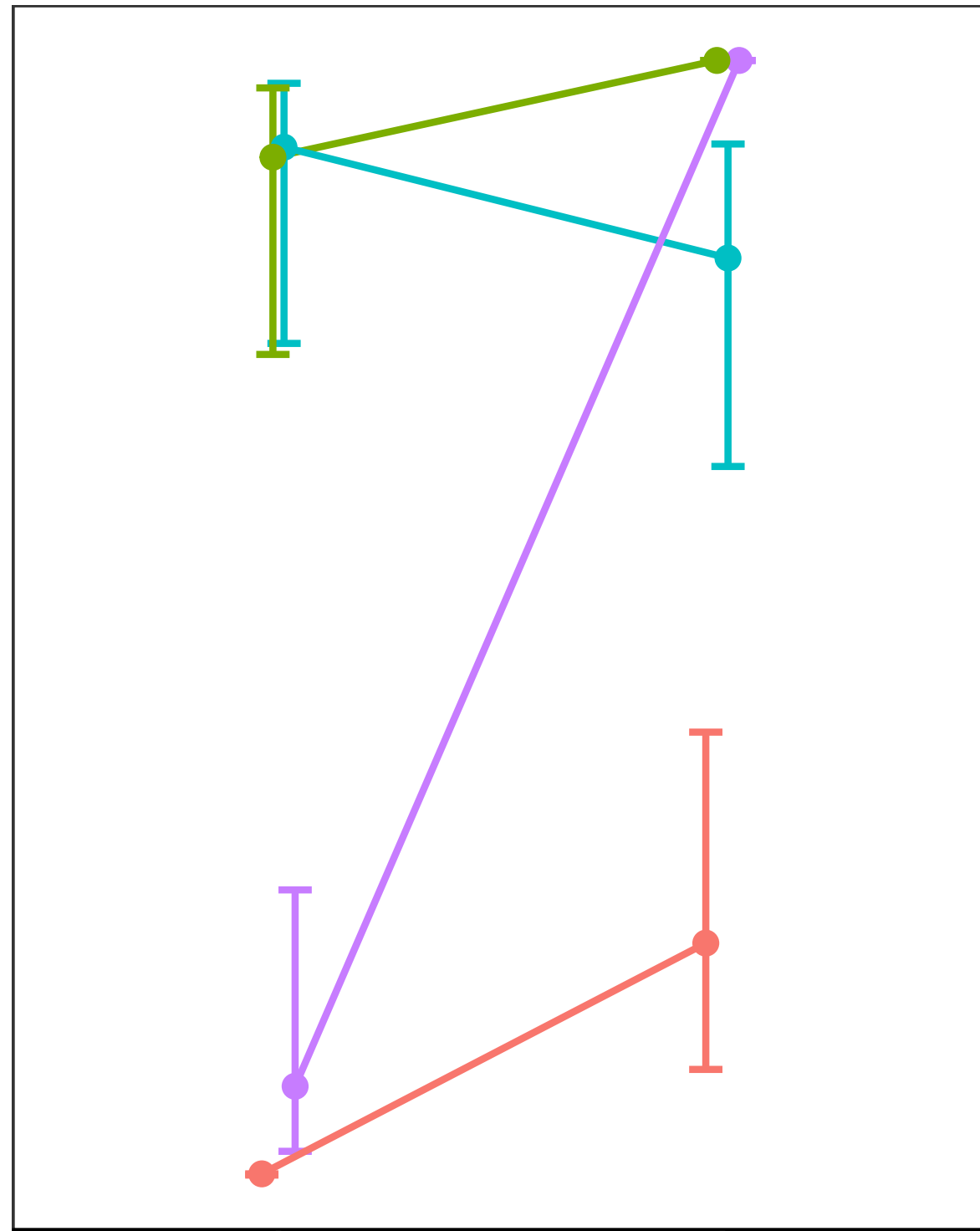
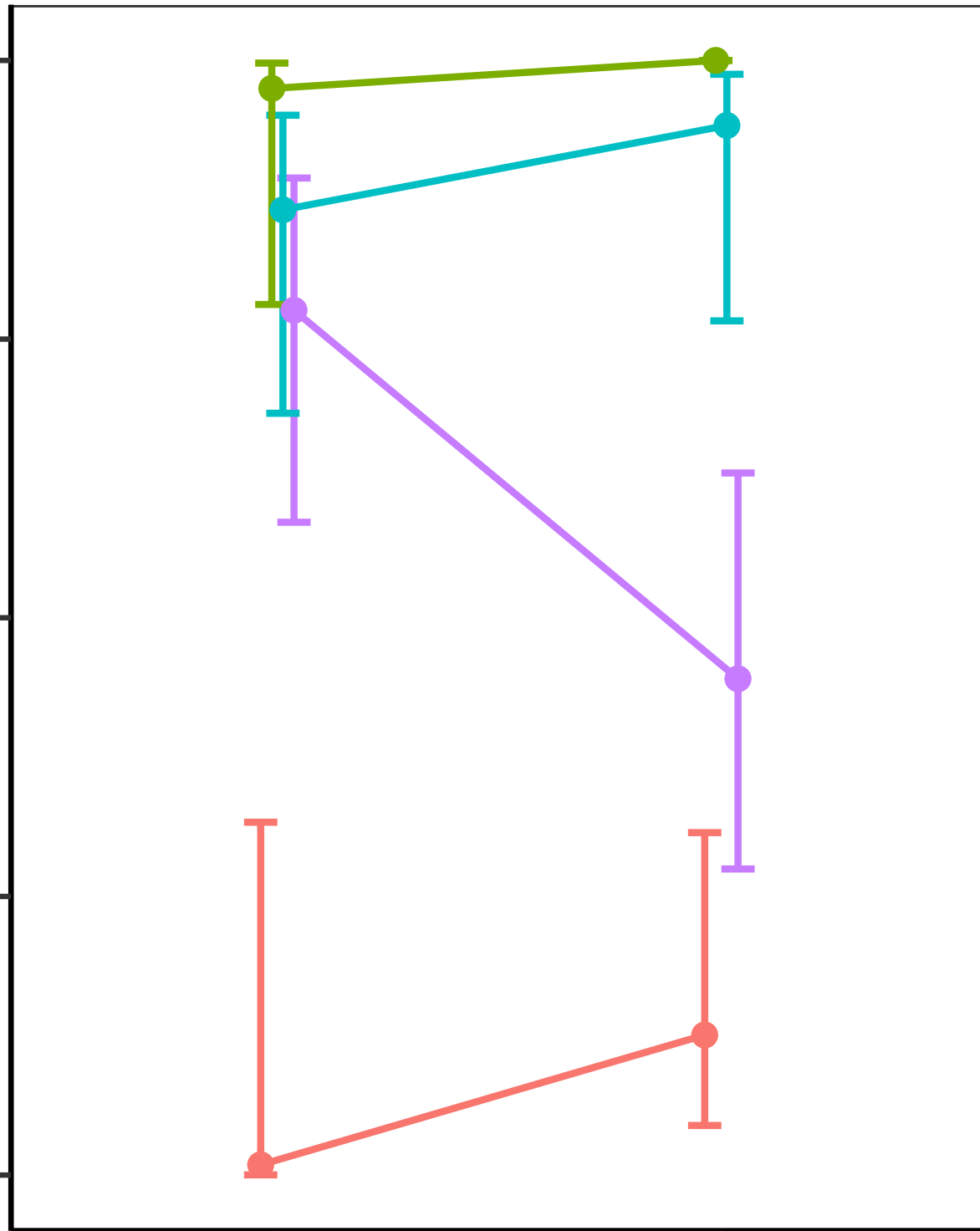
Species

Australian Bass

Hardyheads

Smelt

Silver Perch



Percentage of fish entering

Hardyheads

Smelt

Both

