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1 **Article type: Policy Direction**

2

3 **Seeking convergence on the key concepts in ‘no net loss’**

4 **policy**

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34 **Summary**

- 35 1. Biodiversity conservation policies incorporating a no net loss (NNL) principle are
36 being implemented in many countries. However, there are linguistic and conceptual
37 inconsistencies in the use of terms underlying these NNL policies.
- 38 2. We identify inconsistencies that emerge in the usage of eight key terms and phrases
39 associated with NNL policies: *biodiversity*; *frames of reference* (*i.e. baselines*,
40 *counterfactuals*); *no net loss*; *mitigation hierarchy*; *biodiversity offset*; *in-kind/out-of-*
41 *kind*; *direct/indirect*; and *multipliers*.
- 42 3. For each term, we make recommendations to support conceptual convergence,
43 reduce ambiguity and improve clarity in communication and policy documentation.
44 However, we also warn of the challenges in achieving convergence, especially given
45 the linguistic inconsistencies in several of these key concepts among countries in
46 which NNL policies are employed.
- 47 4. *Policy implications*. The recommendations made in this article, on improving clarity
48 and supporting convergence on key no net loss (NNL) concepts, should help
49 eliminate ambiguity in policy documentation. This is crucial if policymakers are to
50 design robust policies that are: (i) transparent; (ii) translatable into practice in a
51 consistent manner; and, (iii) sufficiently understood and supported by stakeholders to
52 be effective in practice.

53

54 **Keywords:** biodiversity offset, compensation, conservation, counterfactual, frame of
55 reference, mitigation hierarchy, multiplier, no net loss, policy terminology

56

57 **Introduction**

58 The principle of 'no net loss' (NNL) of biodiversity has been embraced by governments
59 (Madsen et al., 2011), multinational corporations, and financial institutions such as the
60 International Finance Corporation (IFC, 2012; Rainey et al., 2014). In most contexts, NNL
61 requires that biodiversity losses associated with development are quantified and any
62 unavoidable impacts fully compensated for by commensurate gains. NNL is generally
63 associated with a 'mitigation hierarchy', under which project developers seek NNL by
64 sequentially avoiding, minimizing, restoring, and offsetting any predicted impacts (Gardner et
65 al., 2013). Associated mechanisms, e.g. biodiversity banking, have become prominent
66 components of the conservation toolkit.

67

68 Despite the proliferating literature on NNL, particularly offsetting (Calvet et al., 2015), the lack
69 of convergence on the usage of key terms is contributing to significant conceptual confusion.
70 For instance, what is known as biodiversity offsetting in some regions (e.g. Australia, UK) is
71 labelled compensatory mitigation elsewhere (e.g. US; Madsen et al., 2011; Box 1). Certain
72 biodiversity offsets in Germany (*Ausgleichsmaßnahmen* or 'compensation offsets') could
73 potentially be interpreted as restoration measures (i.e. a different stage in the mitigation
74 hierarchy) (Tucker et al., 2014). NNL can also be evaluated in various ways resulting in
75 different perceptions as to what 'no net loss' implies (see Bull et al., 2014a). For example,
76 Pickett et al. (2013) discuss a fixed pre-development baseline for evaluating offsets at the
77 Sydney Olympic Park development. But dynamic baselines are also sometimes employed –
78 such as for the Oyu Tolgoi mine in Mongolia (TBC & FFI, 2012), where on-going background
79 habitat deterioration rates were used in determining net outcomes. This potential for
80 confusion is compounded by the fact that NNL-type policies are being developed and
81 implemented across the world in a variety of different languages, which do not necessarily
82 have terms that directly correlate (Table 1).

83

84 *'No net loss' in different languages*

85 Modern NNL policy, incorporating what is today called 'offsetting', grew out of national
86 legislation in the 1970s in both the US (where offsetting is known as 'compensatory
87 mitigation') and several other countries such as Germany (which distinguishes between
88 'compensation restoration' and 'substitution restoration', both of which could potentially be
89 considered 'offsets') and France (Madsen et al., 2011; Tucker et al., 2014).

90

91 The use of the term 'offset' for biodiversity (lagging behind the emergence of carbon offsetting
92 as a concept) seems to have originated more recently via the emergence of the Business and
93 Biodiversity Offsets Programme in 2004 and Australian policies throughout the 2000s
94 (Madsen et al., 2011; Maron et al., 2015).

95

96 Vagueness around terms can also arise in NNL policy development as a result of linguistic
97 uncertainty, during translation of key concepts between different languages (ten Kate &
98 Crowe, 2014). Here, we illustrate this using the example of the various terms used for
99 “biodiversity offset”. Translation of that word can result in conflation of the terms
100 compensation, mitigation, offset and so on (Table 1).

101

102 Terminological confusion can lead to misunderstandings about what NNL policy should, or is
103 designed to, achieve (Gordon et al., 2015), in turn fuelling escalation of debates over the
104 validity of the approach (e.g. Apostolopoulou & Adams, 2015). NNL is the focus of much
105 environmental policy development, for example by the International Union for the
106 Conservation of Nature (IUCN, 2016) and the EU (Tucker et al., 2014). It is thus imperative
107 that key concepts underpinning NNL are clarified and understood in a consistent way, by all
108 stakeholders involved in policy development and project implementation. If not, there is a risk
109 that nascent NNL policies and influential guidance will incorporate vague or misguided
110 concepts that are open to misinterpretation, potentially weakening conservation outcomes.

111

112 Informed by our involvement in reviews of the outstanding challenges for NNL (Bull et al.,
113 2013; Maron et al., 2016), and by concepts which in our experience are most commonly
114 misinterpreted in practice, we identified eight terms underpinning NNL policy that have been
115 used inconsistently in the literature (including, admittedly, by the authors). For each, we briefly
116 discuss the importance of the term, and potential implications of semantic and conceptual
117 inconsistencies. Then, we attempt to provide clarity around the concepts to which the eight
118 terms refer, in the context of NNL policy.

119

120 **Key ‘no net loss’ concepts**

121 *1. Biodiversity*

122 NNL is generally framed as managing and trading losses and gains of biodiversity, so it is
123 important to define what NNL policies mean when referring to ‘biodiversity’. The Convention
124 on Biological Diversity (CBD) recognizes biodiversity at genetic, species and ecosystem
125 levels, and that diversity is driven through complex relationships between biotic and abiotic
126 components and the variability within them (CBD, 2015). Yet the CBD definition of biodiversity
127 – while accepted by many current NNL guidelines (e.g. BBOP, 2012; IFC, 2012) – is evidently
128 not what is intended when NNL objectives refer to ‘biodiversity’. Achievement of NNL for
129 biodiversity using the CBD definition is not only practically impossible to demonstrate, but
130 impossible in principle – e.g. biota carry unique genetic combinations, so exact replacement is
131 not possible.

132

133 Under NNL policies, it is standard either to try and use surrogates for total biodiversity, or a
134 specific set of biological targets that are of interest (e.g. charismatic or threatened species)
135 without claiming that all biota are represented. These measures are mostly species or habitat

136 based, sometimes incorporating processes (e.g. US wetland banking) but rarely considering
137 genes (Bull et al., 2014b). Whichever measures are used as targets in NNL policy, it is only
138 for those targets that the policy is designed to achieve neutral outcomes. Even then, if the
139 measure is a composite of multiple biological characteristics (e.g. condition and area), the
140 potential for substitution means the degree to which neutral outcomes are achieved for each
141 component of the composite measure is not certain (e.g. McCarthy et al., 2004). Yet
142 continuing to use the all-encompassing term 'biodiversity', with its established meaning,
143 implies otherwise.

144

145 We encourage greater efforts by those implementing NNL to clearly state which elements of
146 biodiversity are actually incorporated – with no policy claiming NNL of 'biodiversity' more
147 widely. For example, the stated aim of the "Net Positive Impact" biodiversity strategy for the
148 Oyu Tolgoi project in Mongolia is to achieve "Net Positive Impact (NPI) or No Net Loss (NNL)
149 on biodiversity" up to 2036 (TBC & FFI, 2012). However, the focus for Oyu Tolgoi is in reality
150 the subset of biodiversity features for which NNL or a NPI would be required under PS6 (IFC,
151 2012): one plant, 15 vertebrates, and five habitat types. Thus, an accurate claim for the
152 project would be that the strategy targets NPI for key biodiversity features identified (not
153 biodiversity in general).

154

155 *2. Frames of reference, baselines and counterfactuals*

156 Fundamental to achieving NNL is the frame of reference against which it is evaluated. 'Frame
157 of reference' can be considered an umbrella term for any reference state – including, but not
158 limited to, baselines, scenarios and counterfactuals (Bull et al., 2014a). These terms are often
159 conflated.

160

161 The term 'baseline' has various meanings even within the NNL literature (Maron et al., 2015).
162 For example, 'baseline' may refer to fixed conditions, such as the current state of a system, or
163 a past reference state. A baseline can also refer to a dynamic scenario, reflecting on-going
164 rates of background change, such as the estimated trends for a biodiversity surrogate in the
165 absence of NNL policy (Bull et al. 2014a). Counterfactuals are scenarios capturing what
166 would have occurred under different circumstances, but as they represent a version of reality
167 that is never realized, they can only ever be estimated, and multiple counterfactuals may be
168 plausible. Counterfactuals are necessary in order to attribute additionality, that is, the 'impact'
169 or difference a set of actions made, relative to what was likely to have occurred otherwise
170 (Ferraro & Pattanayak, 2006).

171

172 While reference frames are usually set by regulatory requirements, this information is
173 sometimes not clearly articulated and only implicit. We have previously called for
174 policymakers to be more explicit in specifying the frames of reference being assumed (Bull et
175 al., 2014a; Gordon et al., 2015; Maron et al., 2015). In general, NNL policy remains weighted

176 towards implicitly using current system states (e.g. German policy; Tucker et al., 2014), or
177 counterfactuals of substantial decline (e.g. Australian policies; Maron et al., 2015). We
178 suggest that the term baseline be appropriately modified whenever used, in order to specify
179 the type of reference frame to which it refers. For example, a baseline representing the state
180 of a system immediately prior to development is a 'fixed pre-development baseline'.
181 Conversely, baselines against which gains and losses are to be evaluated could be referred
182 to as 'crediting baselines' and 'debiting baselines' respectively, borrowing from the carbon
183 literature (Maron et al., 2015). The relationship between baselines and counterfactuals
184 requires care – by definition, a counterfactual is never actually observed or measured,
185 whereas a baseline often is. However, observed baselines can form the basis for developing
186 counterfactuals.

187

188 3. *No net loss*

189 Without specification of target ecosystem components and an appropriate frame of reference,
190 NNL could mean different things depending upon interpretation. It is easy to see how a policy
191 objective of 'no net loss of biodiversity or better' (BBOP, 2012) could be presumed by the
192 non-specialist to: (a) apply to all biodiversity; and, (b) be measured against a fixed current
193 baseline. NNL could be thus be interpreted to result in improvement over time for regional
194 biodiversity, compared to the current situation. This perception is sometimes reinforced by
195 policymakers, e.g. in the UK ("[biodiversity offsets offer] an exciting opportunity to look at how
196 we can improve the environment as well as grow the economy"; see Gordon et al., 2015).

197

198 Yet these characteristics (a, b) are not usually intended, and in isolation, the NNL policy
199 principle does not generally result in gains for conservation. Realising this might help lower
200 stakeholder expectations to realistic levels, mitigate concerns that NNL is simply
201 greenwashing, and avoid offsetting being mistakenly presented as an opportunity to 'improve
202 the environment'. Where NNL policy contains an additional requirement for Net Gain, as is
203 sometimes required (e.g. IFC, 2012), then this still does not necessarily mean an absolute
204 decline in biodiversity is avoided – depending upon the frame of reference from which gains
205 are measured. For example, if the counterfactual for a region involved a particularly steep on-
206 going background decline, then Net Gain could technically be achieved by establishing a
207 shallower rate of decline in the region, even if the development and associated offsets
208 allowed a decline to continue in absolute terms (Gordon et al., 2011). Further, there are
209 fundamental differences between NNL and NG as policy principles – they represent different
210 underlying conservation philosophies, encourage different stakeholder expectations, and may
211 involve different treatments of uncertainty and reference frames – which are not always widely
212 recognised (Bull & Brownlie, 2015).

213

214 We propose that the phrase 'no net loss' is always extended to specify the frame of reference
215 against which NNL is to be achieved. In addition, claiming that NNL policy supports overall
216 environmental improvement should be avoided in most cases.

217

218 4. *Mitigation hierarchy*

219 The implementation of NNL policy ostensibly involves following some mitigation hierarchy.

220 Here, we consider the common categorisation of the mitigation hierarchy: Avoid, Minimize,

221 Restore, Offset (Gardner et al., 2013). That is, predicted biodiversity impacts on projects

222 subject to a NNL requirement should first be avoided through design, then minimized in

223 implementation, then remediated where possible, and finally, any residual impacts

224 compensated for via offsets.

225

226 Putting aside the practical challenges facing implementation of the mitigation hierarchy (Bull

227 et al., 2013), a key conceptual challenge is the linguistic vagueness in the way the hierarchy

228 is specified. This results in problems: (i) it is not always clear whether an action represents an

229 avoidance or minimization measure (e.g. carrying out construction works outside of the

230 breeding season for protected fauna); (ii) the third category of the hierarchy is subject to some

231 variety in language, being alternatively labelled 'rehabilitation', 'remediation' (BBOP, 2012),

232 and 'restoration' (IFC, 2012), and all three terms are conflated; (iii) it is unclear at what point

233 restoration activities stop being part of the third stage of the hierarchy, and become

234 biodiversity offsets; and, (iv) biodiversity offsets are sometimes labelled 'compensatory

235 mitigation', causing confusion with the rest of the mitigation hierarchy.

236

237 Despite work having gone into clarifying such questions (e.g. Ekstrom et al., 2015), points (i –

238 iv) above require additional exploration and clarification. We suggest that an avoidance

239 measure is one which, once designed into the project, requires no further action to eliminate

240 the corresponding impacts (e.g. choosing not to extract minerals on a site so as to leave

241 important habitat untouched), whereas minimization measures require on-going action to

242 eliminate corresponding impacts (e.g. carrying out extraction activities during certain times of

243 year so as to avoid the nesting season of a bird species). Both are preventative actions,

244 whereas restoration and offsetting are compensatory actions.

245

246 We argue that the third category of the mitigation hierarchy should be labelled 'remediation',

247 because actions in this category specifically relate to reversing impacts caused by the

248 development to which the hierarchy is being applied. Remediation, by definition, involves

249 reversing damages that one has caused (e.g. replanting an area of vegetation that was

250 cleared to allow construction access). Restoration and rehabilitation, conversely, refer to more

251 general processes ("Rehabilitation emphasizes the reparation of ecosystem processes,

252 productivity and services whereas the goals of restoration **also** include the re-establishment

253 of the pre-existing biotic integrity in terms of species composition and community structure”;
254 SER, 2004).

255

256 Whilst remediation may involve ecological restoration, it is different to biodiversity offsets.
257 Offsets do not reverse damages, they compensate for damages in some other way (e.g.
258 planting a new area of vegetation to compensate for project-related clearances). This
259 distinction can be illustrated as a difference between *Ausgleichsmaßnahmen* (which might be
260 interpreted as remediation) and *Ersatzmaßnahmen* (which might be interpreted as offsetting)
261 measures under German NNL (Tucker et al., 2014).

262

263 Finally, despite the widespread use of the term, we discourage describing biodiversity offsets
264 as ‘compensatory mitigation’. Compensation is a term that applies to a broader class of
265 measures than offsets (Bull et al., 2013), and ‘compensatory mitigation’ could equally be used
266 to describe the third stage of the hierarchy (‘remediation’).

267

268 5. *Offset*

269 The word ‘offset’ means to counteract something by having an equal and opposite force or
270 effect (Oxford Dictionary of English). An offset exchange requires that the ecological targets –
271 such as particular species or habitats – are not diminished in net terms compared to what
272 would have occurred without the impact and offset (Maron et al., 2012).

273

274 The more general terms ‘compensate’ and ‘mitigate’ are often used interchangeably with
275 ‘offsets’ (Madsen et al., 2011). This is problematic as it creates confusion about what
276 constitutes an offset, and where the bar lies for achieving true NNL. For example in the US,
277 banks of created or restored wetlands – effectively supplying true biodiversity offset credits –
278 are labelled ‘mitigation banks’. Conversely, many projects seek to offer financial
279 compensation, education schemes, or research and monitoring funds as part of ‘offset
280 packages’ (e.g. Oyu Tolgoi; TBC & FFI, 2012). We argue that such activities do not constitute
281 true biodiversity offsets unless measurable and commensurate gains in the biota targeted are
282 achieved through these mechanisms, but the distinction can be hard to make. Furthermore, to
283 qualify as an offset, there must be demonstrably quantifiable equivalence between what is
284 lost and gained, and the term offset should be quarantined for this use only. An offset can
285 therefore be seen as a specific and rigorously quantified type of compensation measure. We
286 recommend the broader term ‘compensation’ be reserved for other types of actions that do
287 not meet our definition of an offset.

288

289 6. *In-kind versus out-of-kind*

290 Biodiversity offsets are often categorised as ‘in-kind’ or ‘out-of-kind’. These terms refer to the
291 biodiversity attributes being impacted and offset, and whether they are similar or different,
292 respectively. Note, in-kind is not the same as ‘on-site’ (on-site offsets can be either in-kind or

293 out-of-kind, as can off-site offsets), even though the terms are sometimes used
294 interchangeably. Under the CBD definition of biodiversity, all offsets are out-of-kind, as
295 biodiversity in any two places can never be truly identical. However, since specific surrogates
296 of biodiversity are the targets under NNL policy, in-kind offsets are possible with respect to
297 these surrogates. An important question is whether out-of-kind trades with respect to the
298 surrogates can ever qualify as true offsets.

299
300 One widely accepted type of out-of-kind offsetting is referred to “trading up” (BBOP, 2012),
301 where offsets seek gains in components of biodiversity of higher conservation value than
302 those impacted. For example, impacts on a common and unthreatened ecological community
303 (e.g. fallow agricultural land) being offset by gains for a more threatened community (e.g.
304 wetland). More generally, so-called ‘strategic offsetting’ has been advocated as an effective
305 approach (Sochi & Kiesecker 2016), integrating offsetting with conservation planning. This
306 makes use of well-developed techniques for prioritising locations for conservation activities
307 based upon factors such as complementarity, irreplaceability, species rarity, cost, and threat.
308 Out-of-kind offsets might sometimes, under such an approach, achieve benefits more valued
309 from a conservation perspective compared to strict like-for-like offsetting (Habib et al., 2013;
310 Bull et al., 2015). The downside of this approach is that it either removes the clear connection
311 between losses and gains, or obscures the targets of the exchange.

312
313 We argue that out-of-kind exchanges of biodiversity (including trading up) should not be
314 referred to as ‘offsets’ in the strict sense unless the biodiversity surrogates upon which the
315 policy operates are specifically designed to be fungible. For instance, where Habib et al.
316 (2013) propose the use of caribou conservation as a flexible offset for vegetation clearances
317 in western Canada, this would be labelled strategic compensation – whilst measures that
318 compensated like-for-like with habitat restoration would be true offsets. In proposing this
319 position, we again emphasize that true fungibility does not exist for trades in actual
320 components of biodiversity e.g. individual organisms (Salzman & Ruhl, 2000), so in practice
321 ‘in-kind’ means ‘fungible in relation to the specified biodiversity metric’. Equally, we
322 acknowledge a practicality – developers may be less likely to attempt strategic compensation
323 measures, involving gains of very high conservation value, if they are strictly required by
324 policy to demonstrate that they have implemented some kind of ‘offset’, meaning potentially
325 foregone opportunities for substantial conservation gains. However without making a
326 distinction of this sort between in-kind and out-of-kind trades, we risk the outcomes from
327 offset activities becoming so varied and ambiguous that the fundamental NNL principle
328 becomes meaningless.

329
330 *7. Direct versus indirect offsets*

331 The terms ‘direct’ and ‘indirect’ reflect multiple conceptual dichotomies, variously being used
332 to distinguish between offsets on the basis of: (a) biodiversity outcomes; (b) type of action

333 undertaken (e.g. restoration, protection); and, (c) mechanisms through which the offset is
334 delivered (Miller et al. 2015). For an example of (a), the terms are commonly used to
335 distinguish between actions with direct, measurable benefits for target biota (e.g. protection or
336 enhancement of habitat) from those without (e.g. public education). But the dichotomy has
337 also been made on the basis of (b) whether offset actions involved purchasing land, or
338 addressing threats to species in an alternative way. Alternatively, the distinction is based not
339 on offset outcomes, but on (c) the pathway for delivery – i.e. direct offsets are provided or
340 purchased by the proponent of the impact, while indirect offsets involve payment to a third
341 party (such as a government) who assumes liability for finding an offset to compensate for
342 losses. To further confuse matters, the impacts giving rise to offsets can be either direct or
343 indirect (Curran et al., 2015), with a comparably inconsistent use of the terms – although in
344 this article we focus on the application of the terms to offsetting.

345

346 We recommend the direct/indirect dichotomy be reserved for category (c) above, i.e. the
347 pathway through which offsets generate measurable benefits for target biota. Activities that do
348 not achieve such an outcome should not be defined as offsets (see point 5). Often, funding for
349 research and increasing community awareness would fall into the ‘indirect’ category by this
350 definition, but not necessarily in all cases – for instance, Weston et al. (2011) describe
351 measurable benefits to shorebird nesting success being directly attributed to increased
352 signage and community education. Similarly, indirect pathways of funding for an offset can
353 still, in theory, generate a direct benefit for the target biota, although greater risks may be
354 involved.

355

356 8. *Multipliers*

357 Under>NNL, a ‘multiplier’ can refer to the relative quantity of biodiversity gained and lost at
358 offset and impact sites respectively, or the relative areas over which the impact and the offset
359 actions are undertaken. So for example, a multiplier of two implies that the gains from the
360 offset were required to be twice as large as losses from the area impacted – or, that offsets
361 occupying twice the area of the impacts would be expected to generate a gain equivalent to
362 the losses. The term ‘compensation ratio’ is also commonly used to refer to the relationship
363 between gains and losses (Laitila et al., 2014). Multipliers are often not labelled as such in
364 offset implementation, or simply not specified at all (Bull et al., in review).

365

366 Multipliers are one strategy amongst many (e.g. equivalency analysis; Quétier & Lavorel,
367 2012) for managing uncertainties in biodiversity gains from offset activities, and to account for
368 time lags in which the offset gains accrue compared to impacts (through time discounting, e.g.
369 Gibbons et al., 2015). They are also used for other reasons – e.g. imposing higher
370 requirements on offsets for threatened habitats (South Africa; Laitila et al., 2014). In practice,
371 multipliers are often determined based upon negotiation between stakeholders involved in a
372 given development, rather than as a result of robust scientific considerations (Bull et al., in

373 review). Multipliers are sometimes less than or equal to one (i.e. biodiversity gains are smaller
374 than losses in terms of the biodiversity measure specified; e.g. Quigley & Harper, 2006). In
375 such cases the term 'multiplier' is appropriate, but subject to the considerations of appropriate
376 baselines (see point 2), the trade should not necessarily be treated as achieving NNL.

377

378 We recommend that multipliers or compensation ratios and their purpose be explicitly
379 specified with justification in NNL policies and projects. For example, whether the goal of a
380 multiplier is to increase the amount of benefit expected from an offset to achieve an outcome
381 of better than NNL, or whether it is to adjust for factors such as uncertainty and time lags,
382 should be clearly specified. Further, any multipliers less than one ought to be particularly
383 closely scrutinised. It must be more widely recognised that multipliers that account for
384 scientific matters such as uncertainty and time lags are a crucial component of achieving
385 NNL, and therefore not necessarily open to negotiation when the goal is NNL.

386

387 **Concluding remarks**

388 As can be seen, there remains considerable linguistic inconsistency around NNL policies,
389 arising from both vagueness in the terms themselves and from the variation in standard
390 regulatory language across jurisdictions. In our experience, this causes considerable
391 conceptual confusion. Here, we have highlighted eight key terms associated with NNL
392 policies that have yet to achieve linguistic and conceptual convergence – making suggestions
393 as to how such convergence might be sought (Table 2). We do not claim that these are the
394 only NNL terms applied inconsistently, but they are some of the most fundamental and
395 therefore important to clarify. The terms cover interrelated aspects across NNL policy (Fig. 1),
396 and so the vagueness that arises in each is compounded. We accept that the language of
397 policy and regulations varies across jurisdictions, and that linguistic uncertainty arises when
398 translating terms between spoken languages, and accordingly our intention in writing this
399 article is not to encourage changes in the terms employed by existing guidelines or
400 legislation. Rather, it is to seek shared understanding of the concepts underlying the NNL
401 principle, whatever language is then used to express those concepts. We consider it unlikely
402 that all researchers and practitioners will agree with our suggestions here, but welcome any
403 discussion that our proposals encourage on this crucial topic.

404

405 **Data accessibility**

406 Data have not been archived because this article does not contain data.

407

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418 **References**

- 419 Apostolopoulou, E. & Adams, W.M. (2015) Biodiversity offsetting and conservation: reframing
420 nature to save it. *Oryx*, DOI: 10.1017/S0030605315000782.
421
- 422 BBOP (Business and Biodiversity Offsets Programme). (2012). Standard on Biodiversity
423 Offsets. BBOP; Washington DC, USA.
424
- 425 Bull, J.W., Suttle, K.B., Gordon, A., Singh, N.J., Milner-Gulland, E.J. (2013) Biodiversity
426 offsets in theory and practice. *Oryx*, 47(3): 369-380.
427
- 428 Bull, J.W., Gordon, A., Law, E., Suttle, K.B. & Milner-Gulland, E.J. (2014a) The importance of
429 baseline specification in evaluating conservation interventions and achieving no net
430 loss of biodiversity. *Conservation Biology*, 28(3): 799–809.
431
- 432 Bull, J.W., Milner-Gulland, E.J., Suttle, K.B. & Singh N.J. (2014b) Comparing biodiversity
433 offset calculation methods with a case study in Uzbekistan. *Biological Conservation*,
434 178: 2–10.
435
- 436 Bull, J.W. & Brownlie, S. (2015) The transition from No Net Loss to a Net Gain of biodiversity
437 is far from trivial. *Oryx*, DOI:10.1017/S0030605315000861.
438
- 439 Bull, J.W., Hardy, M.J., Moilanen, A. & Gordon, A. (2015) Categories of flexibility in
440 biodiversity offsetting, and the implications of out-of-kind ecological compensation.
441 *Biological Conservation*. DOI:10.1016/j.biocon.2015.08.003.
442
- 443 Bull, J.W., Lloyd, S. & Strange, N. (in review) Multipliers used in practice under No Net Loss
444 policies. *Conservation Letters*.
445
- 446 Calvet, C., Guillaume, O. & Claude N. (2015) Tracking the origins and development of
447 biodiversity offsetting in academic research and its implications for conservation: A
448 review. *Biological Conservation*, doi: 10.1016/j.biocon.2015.08.036.
449
- 450 CBD (Convention on Biological Diversity) (2015) Article 2: Use of Terms [available at:
451 <https://www.cbd.int/convention/articles/default.shtml?a=cbd-02>] (accessed December
452 2015).
453
- 454 Curran, M., Hellweg, S. & Beck, J. (2015) The jury is still out on biodiversity offsets: reply to
455 Quétier et al. *Ecological Applications*, 25(6):1741-1746.
456

- 457 Ekstrom, J., Bennun, L. & Mitchell, R. (2015) A cross sector guide for implementing the
458 mitigation hierarchy. Cross Sector Biodiversity Initiative [available at:
459 [http://www.ipieca.org/publication/cross-sector-guide-implementing-mitigation-
461 hierarchy](http://www.ipieca.org/publication/cross-sector-guide-implementing-mitigation-
460 hierarchy)].
- 462 Ferraro, P.J. & Pattanayak, S.K. (2006) Money for nothing? A call for empirical evaluation of
463 Biodiversity Conservation Investments. *PLoS Biology*, 4(4):482-488.
464
- 465 Gardner, T.A., von Hase, A., Brownlie, S., Ekstrom, J.M.M., Pilgrim, J.D., *et al.* (2013)
466 Biodiversity offsets and the challenge of achieving no net loss. *Conservation Biology*,
467 DOI: 10.1111/cobi.12118.
468
- 469 Gibbons P., Evans M., Maron M., Gordon A., Le Roux D., von Hase A., Lindenmayer D.,
470 Possingham H.P. (2015) A loss-gain calculator for biodiversity offsets and the
471 circumstances in which no net loss is feasible. *Conservation Letters*. doi:
472 10.1111/conl.12206.
473
- 474 Gordon, A., Bull, J.W., Wilcox, C. & Maron, M. (2015) Perverse incentives risk undermining
475 biodiversity offset policies. *Journal of Applied Ecology*, 52: 532–537. DOI:
476 10.1111/1365-2664.12398.
477
- 478 Gordon, A., Langford, W.T., Todd, J.A., White, M.D., Mullerworth, D.W, Bekessy, S.A. (2011)
479 Assessing the impacts of biodiversity offset policies. *Environmental Modelling and
480 Software*, 26: 1481–1488. DOI 10.1016/j.envsoft.2011.07.021.
481
- 482 Habib, T.J., Farr, D.R., Schneider, R.R. & Boutin, S. (2013) Economic and Ecological
483 Outcomes of Flexible Biodiversity Offset Systems. *Conservation Biology*. DOI:
484 10.1111/cobi.12098.
485
- 486 IFC (International Finance Corporation) (2012) Performance Standard 6: Biodiversity
487 Conservation and Sustainable Management of Living Natural Resources. IFC, World
488 Bank Group.
489
- 490 IUCN (International Union for the Conservation of Nature) (2016) IUCN policy on biodiversity
491 offsets [available at:
492 http://cmsdata.iucn.org/downloads/iucn_biodiversity_offsets_policy_jan_29_2016.pdf].
493
- 494 Laitila, J., Moilanen A. & Pouzols F. M. (2014) A Method for Calculating Minimum Biodiversity
495 Offset Multipliers Accounting for Time Discounting, Additionality and Permanence.
496 *Methods in Ecology and Evolution*, 5(11):1247–54.

- 497
- 498 Madsen, B., Carroll, N., Kandy, D. & Bennett, G. (2011) State of Biodiversity Markets Report:
499 Offset and Compensation Programs Worldwide. Forest Trends; Washington D.C., USA.
500
- 501 Maron, M., Hobbs, R.J., Moilanen, A., Matthews, J.W., Christie, K., *et al.* (2012) Faustian
502 Bargains? Restoration Realities in the Context of Biodiversity Offset Policies. *Biological*
503 *Conservation*, 155:141–48.
504
- 505 Maron, M., Bull, J.W., Evans, M.C. & Gordon, A. (2015) Locking in loss: baselines of decline
506 in Australian biodiversity offset policies. *Biological Conservation*,
507 doi:10.1016/j.biocon.2015.05.017.
508
- 509 Maron, M., Ives, C., Kujala, H., Bull, J.W., Maseky, F.J.F., *et al.* (2016) Taming a wicked
510 problem: resolving controversies in biodiversity offsetting. *BioScience*,
511 doi:10.1093/biosci/biw038.
512
- 513 McCarthy, M.A., Parris, K.M., van der Ree, R., McDonnell, M.J., Burgman, M.A., *et al.* (2004)
514 The habitat hectares approach to vegetation assessment: An evaluation and
515 suggestions for improvement. *Ecological Management & Restoration*, 5(1):24-27.
516
- 517 Miller, K.L., Trezise, J.A., Kraus, S., Dripps, K., Evans, M.C., *et al.* (2015) The development of
518 the Australian environmental offsets policy: from theory to practice. *Environmental*
519 *Conservation*, 42(4):306–314.
520
- 521 Pickett, E.J., Stockwell, M.P., Bower, D.S., Garnham, J.I., Pollard, *et al.* (2013) Achieving no
522 net loss in habitat offset of a threatened frog required high offset ratio and intensive
523 monitoring. *Biological Conservation*, 157:156-162.
524
- 525 Quétier F., & Lavorel S. (2012) Assessing ecological equivalence in biodiversity offset
526 schemes: key issues and solutions. *Biological Conservation*, 144(12):2991-2999.
527
- 528 Quigley, J.T. & Harper, D.J. (2006) Effectiveness of fish habitat compensation in Canada in
529 achieving no net loss. *Environmental Management*, 37(3):351-366.
530
- 531 Rainey H.J., Pollard, E.H.B., Dutson, G., Ekstrom, J.M.M., Livingstone, S.R., *et al.* (2014) A
532 review of corporate goals of No Net Loss and Net Positive Impact on biodiversity. *Oryx*,
533 doi:10.1017/S0030605313001476.
534
- 535 Salzman, J. & Ruhl, J.B. (2000) Currencies and the commodification of environmental law.
536 *Stanford Law Review*, 53(3):607-694.

- 537
- 538 SER (Society for Ecological Restoration) (2004) Primer on ecological restoration. SER;
539 Washington, D.C., USA [available at: [http://www.ser.org/resources/resources-detail-](http://www.ser.org/resources/resources-detail-view/ser-international-primer-on-ecological-restoration)
540 [view/ser-international-primer-on-ecological-restoration](http://www.ser.org/resources/resources-detail-view/ser-international-primer-on-ecological-restoration)].
- 541
- 542 Sochi, K. & Kiesecker, J. (2016) Optimizing regulatory requirements to aid in the
543 implementation of compensatory mitigation. *Journal of Applied Ecology*, 53(2):317-322.
544
- 545 TBC & FFI (2012) Oyu Tolgoi Net Positive Impact Forecast. Unpublished draft report of The
546 Biodiversity Consultancy Ltd and Fauna & Flora International [available at:
547 http://ot.mn/media/ot/content/page_content/commitments/ESIA/1_ESIA/Biodiversity_Appendices/ESIA_BA5_Net_Positive_Impact_Forecast_for_the_Oyu_Tolgoi_Project.pdf].
548
- 549
- 550 ten Kate, K. & Crowe, M.L.A. (2014). Biodiversity Offsets: Policy options for governments. An
551 input paper for the IUCN Technical Study Group on Biodiversity Offsets. Gland,
552 Switzerland: IUCN. 91pp.
553
- 554 Tucker, G., Allen, B., Conway, M., Dickie, I., Hart, K., *et al.* (2014) Policy Options for an EU
555 No Net Loss Initiative. Report to the European Commission. Institute for European
556 Environmental Policy, London.
557
- 558 Weston, M.A., Ehmke, G.C. & Maguire, G.S. (2011) Nest return times in response to static
559 versus mobile human disturbance. *The Journal of Wildlife Management*, 75(1):252–
560 255.

561 **Table 1.** Terminology used to refer to 'biodiversity offset' type mechanisms in selected
 562 languages
 563

Language	Equivalent terminology for biodiversity offset	English (UK) direct translation	Relevant countries
Chinese (simplified)	shengtai buchang jizhi	eco-compensation mechanism	China
Danish	kompensation	compensation	Denmark
English (Canada)	conservation offset	conservation offset	Canada
English (UK)	biodiversity offset	biodiversity offset	Australia, New Zealand, South Africa, UK
English (US)	compensatory mitigation	compensatory mitigation	US
French	mesures de compensation; compensation écologique	compensation measures; ecological compensation	Canada, France, Madagascar
German	Ausgleichsmaßnahmen; Ersatzmaßnahmen	compensation measures; substitution measures	Germany
Japanese	'satoyama (里山) banking'	[<i>satoyama</i> is the term for a semi-agricultural ecosystem type in Japan]	Japan
Portuguese (Brazilian)	cota de reserva ambiental	environmental reserve certificate	Brazil
Russian	биоразнообразия компенсация	biodiversity compensation	Kazakhstan, Russia, Uzbekistan
Spanish	compensaciones de biodiversidad; medidas compensatorias	biodiversity compensation; compensatory measures	Argentina, Chile, Colombia, Mexico, Peru, Spain, Venezuela
Swedish	ersättning; ekologisk compensation; miljökompensation	compensation / substitution; ecological compensation; environmental compensation	Sweden

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565 **Table 2.** Summary of key recommendations made in this article

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Term	Recommendations
1. Biodiversity	<ul style="list-style-type: none"> • Explain and communicate that biodiversity in NNL policies is not 'total biodiversity' (i.e. CBD definition) • Explicitly state which components of total biodiversity are within scope
2. Frames of reference	<ul style="list-style-type: none"> • NNL always evaluated against some 'frame of reference' • Specify whether frame of reference is a fixed point or trend • Modify the term baseline when it is used, to be more explicit (e.g. 'crediting baseline') • Baselines and counterfactuals are both reference states/trends used for evaluating change, but counterfactuals are (by definition) scenarios that did not actually occur, whereas baselines often do
3. No net loss	<ul style="list-style-type: none"> • Clarify when the goal of NNL policies is not to prevent absolute biodiversity declines • Distinguish clearly between NNL and Net Gain policies
4. Mitigation hierarchy	<ul style="list-style-type: none"> • Develop a more concrete distinction between 'avoidance' and 'minimization' • Ensure that options to forgo development or resource use are considered before any compensatory actions are suggested • Label the third stage of the mitigation hierarchy 'remediation' • Develop a concrete distinction between 'remediation' and 'offset' • Do not label biodiversity offset measures as 'compensatory mitigation'
5. Offset	<ul style="list-style-type: none"> • Do not use 'offsetting' as a label for broader 'compensation measures' which do not meet the stricter definition of offsets • Do not include financial payments within offset packages unless biodiversity gains from those payments are directly quantifiable • Do not include any other interventions with non-quantifiable biodiversity outcomes (e.g. research, education*) in offset packages
6. In-kind/out-of-kind	<ul style="list-style-type: none"> • Seek a new label for out-of-kind offsets, communicating that they are not strictly true offsets
7. Direct/indirect	<ul style="list-style-type: none"> • Reserve 'direct' and 'indirect' to distinguish between the pathway for delivery of an offset, rather than biodiversity outcomes
8. Multipliers	<ul style="list-style-type: none"> • Always specify multipliers, and their intended function • Seek to use multipliers greater than one

* in rare cases, quantifiable ecological benefits may be achievable through education

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569 **Figure 1.** Schematic diagrams for terms discussed in this article. (a) Flow diagram containing
570 all eight terms (grey boxes), and their interrelations. (b) Illustration of key NNL concepts,
571 representing a deteriorating ecosystem as a car driving down a slope. Development impacts
572 increase the steepness of the slope, measures implemented under the mitigation hierarchy
573 return it to the original gradient. NNL is achieved for some components of the ecosystem
574 ('biodiversity') against the frame of reference for an observer in the moving car. Indirect and
575 out-of-kind offsets in this representation might constitute changing the slope of an alternative
576 road. 'Biodiversity' image: modified from <http://www.thebluedotpost.com/> (2014).

